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Postfire management impacts on soil hydrology

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Post-fire management impacts on soil hydrology

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Abstract

Research about soil hydrology after wildfire has widely investigated the impacts of many post-fire management strategies on ecosystems with different characteristics. However, despite this ample literature, clear guidelines about the effectiveness and feasibility of the different restoration techniques in environmental contexts showing variable responses still lack. Furthermore, post-fire hydrological modelling is based on mere adaptations of existing models, which often fail to simulate with accuracy the changes in soil hydrology after fire. After a short review about the effects of wildfire on hydrological processes, this study aims to propose an updated overview of the existing post-fire management techniques at both hillslope (afforestation and seeding, mulching, salvage logging, erosion barriers, soil preparation and other novel techniques) and channel (check dams) scales. Moreover, the results of the most recent studies analysing the feasibility of common hydrological models in predicting runoff and soil erosion are analyzed. Most studies have demonstrated the effectiveness of post-fire management techniques, but some uncertainty remains regarding the opportunity of natural recovering or implementation of soil and vegetation restoration. The optimal solution in fire-affected areas may be a combination of actions (at hillslope and channel scales), whose effectiveness should be evaluated on the watershed scale. The existing hydrological models should be specifically adapted to burned conditions with reliable simulation of soil changes due to fire. Modelling experiences with focus on the effects of post-fire management actions are needed.

Keywords: infiltration; water repellency; runoff; mulching; soil loss; hydrological models.

1. Background

Wildfire is a natural and anthropogenic agent with a long history of influence on terrestrial and aquatic ecosystems [1]. The wildfire effects, which mainly depends on fire severity, extend in

50 several components, such as soil, vegetation, air, and surface and deep water [2], determining strong
51 changes in the ecosystems affected by fire. Wildfire can have positive impacts on soils, increasing
52 fertility and weathering, particularly in fire-affected areas where mild morphological conditions and
53 rapid vegetation cover limit post-fire erosion [3]. The magnitude of these changes varies according
54 to the pre-fire environmental conditions of the burnt areas, wildfire characteristics, and post-fire
55 weather dynamics and human actions [4,5]. The vegetation burning coupled to the alterations in the
56 physico-chemical properties of soils modify the soil hydrology with possible increases in surface
57 runoff as well as soil erosion and degradation rates. Overland flow and eroded sediments generated
58 on hillslopes easily reach the catchment channels, producing severe hydrological effects compared
59 to unburned areas [6]. These negative impacts may lead to loss of biomass productivity and decline
60 on short- to medium-term soil biodiversity [7] beside geomorphologic changes in the rivers and
61 landscape. Moreover, the wildfire effects may extend in the space and in time. As a matter of fact,
62 increases in flood risk and pollution of downstream water bodies can be recorded outside of the
63 burned area, and the pre-fire conditions of the burned ecosystem may be restored after a period
64 varying from few months to several years. Fire-induced changes on soil hydrology also affects
65 forest ecosystem services, including water resource availability, quality of water bodies, erosion and
66 flood control, and biodiversity maintenance [8].

67 This study carries out an updated overview of the most common post-fire management techniques
68 as well as hydrological models reported in a selection of papers published between 2017 and 2021
69 and selected using relevant keywords from Scopus and Web of Science databases. This overview
70 aims at: (i) understanding the effectiveness of each technique on post-fire hydrology of burned and
71 restored soils across different environmental contexts; (ii) analysing whether the available
72 hydrological models are effective in simulating post-fire hydrology. Finally, scientific literature
73 gaps and future research directions are discussed.

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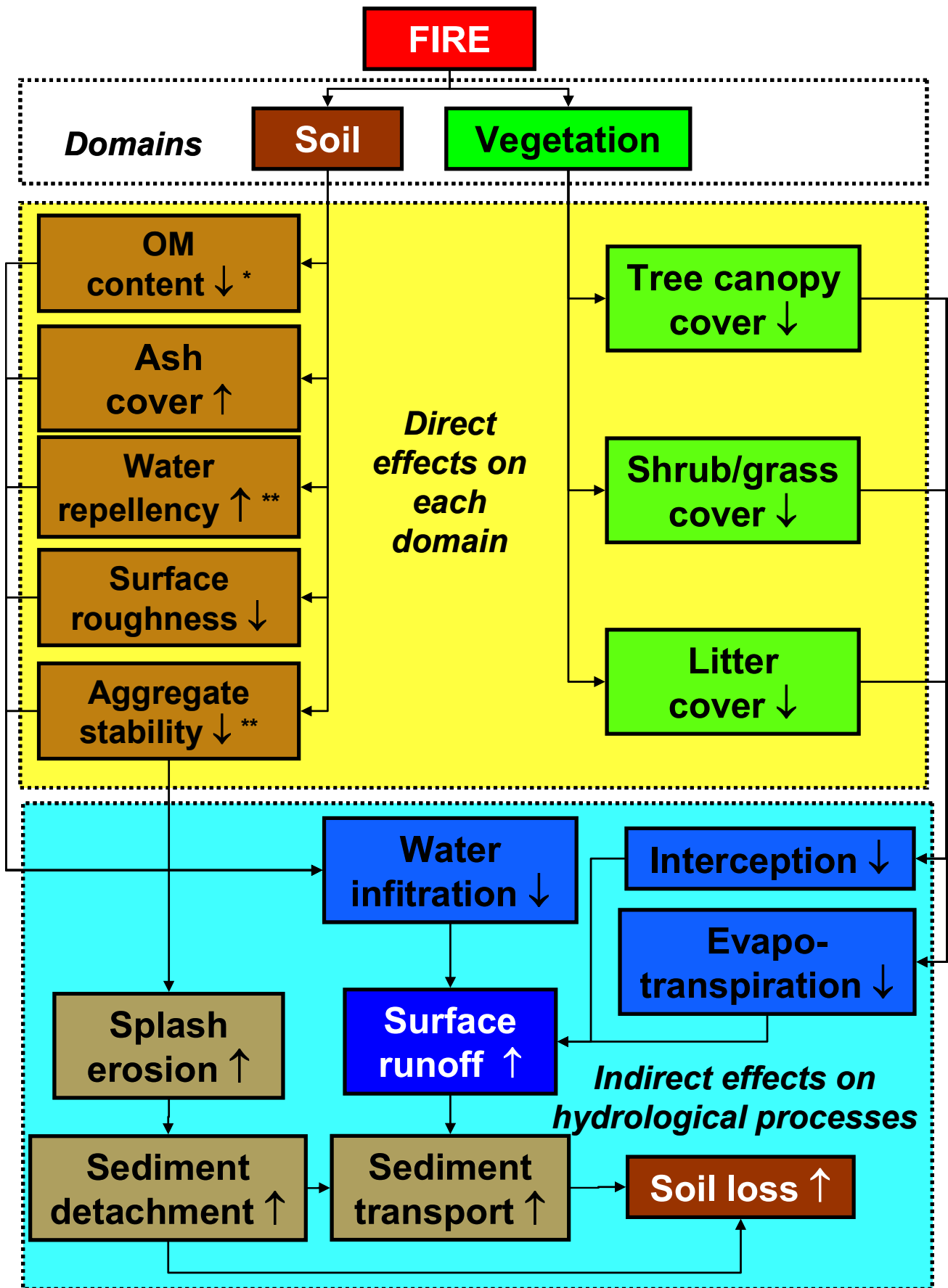
76 **2. Wildfire effects on soil hydrology**

77

78 The occurrence of wildfire produces an increased hydrological response in recently burnt areas,
79 especially during the “window-of-disturbance” [9]. In this period, the soil vulnerability to runoff
80 and erosion effects increases compared the unburned forest areas [4,10]. After the window of
81 disturbance ends, the background pre-fire hydrological conditions tend to be restored throughout
82 periods lasting from few months to several years, that is, after the vegetation cover is re-established
83 and the magnitude of soil erosion is reduced [8,11]. The extent of the fire-induced changes in soil

84 hydrology strictly mainly depends on wildfire intensity (i.e., the energy release rate) and severity
85 (i.e., the magnitude of changes in the burned ecosystem) [12]. However, other environmental
86 factors play important role on changes in fire-affected ecosystems, such as type and physico-
87 chemical properties of soils, topography, fire history, fuel quantity, vegetation species, weather
88 patterns, etc. [13][14]. The wildfire characteristics drive the effects of regenerating vegetation cover
89 on hydrological properties of soil as well as the changes in the properties of the affected soils
90 [15,16]. Fire reduces (in the case of low-severity fire) or completely removes (for fires with high
91 severity) the canopy and ground cover of vegetation and litter. Therefore, interception and evapo-
92 transpiration decreases, and net precipitation increases, leading more water available for runoff
93 [8,11]; moreover, a soil left bare due to vegetation burning becomes more susceptible to raindrop
94 impact and particle detachment. Wildfire also modifies the physico-chemical properties of the soil
95 surface in many ways that influence the hydrological response to precipitation events based on the
96 heat released [17,18] (Figure 1). Wildfire impacts on soil properties can be direct or indirect. Direct
97 impacts, which are related to burning duration and fire temperatures, are usually short and restricted
98 to the upper layer of the soil (few centimetres from the surface). The indirect impacts of wildfire
99 depend on several factors, such as the ash release, vegetation cover, morphology as well as post-fire
100 weather patterns and management [13,19–21]. More specifically, ash is a key driver of the
101 hydraulic response of the burned soils (depending on its depth and type). Ash can either increase the
102 soil water retention and reduce the soil water repellency or might seal the soil surface, reducing
103 water infiltration and increasing surface runoff and flooding [15,18,22]. Ash impact on hydrological
104 characteristics of burned soil depends on its colour; in more detail, black ash, generated by lower
105 fire temperatures, acts as a mulch with a wettable cover for soil, retaining rainwater and improving
106 infiltration, while gray to white ashes of higher severity fires, clog soil pores and generates surface
107 sealing, increasing overland flow and erosion processes [23]. As result of wildfire impacts, in burnt
108 areas, sealing, surface crust formation, pore clogging, and bulk density increase [1]; moreover, soil
109 organic matter and macro-nutrients are lost and its structure can be modified by fire. The depletion
110 of soil organic matter has a substantial effect on soil properties such as the structure as well as
111 chemical and biological properties [19]. In turn, these effects on soil physico-chemical
112 characteristics can influence the hydraulic properties, such as water repellency, water retention,
113 hydraulic conductivity and sorptivity. wildfire particularly impacts on aggregate stability and water
114 repellency of soils, and also these effects on soils depend on temperature and duration [24]. Soil
115 aggregate stability is not altered or slightly increases at temperatures up to 220 °C, while it is
116 strongly reduced between 380 and 460 °C [25]. Soil structure is irreversibly disrupted over 460 °C
117 [24], while, in contrast, clayey soils can show increased aggregate stability at high temperatures

118 [26]. Soil water repellency does not noticeably change at temperatures under 200 °C, increases
119 between 250 and 300 °C, and completely disappears over 300-400 °C [13] [25] (Figure 1).
120 The changes in soil properties due to wildfire determine noticeable impacts on post-fire hydrology,
121 such as reduction in water infiltration or shifts in runoff generation mechanisms [1,15,18].
122 However, the changes in soil hydrology are differentiated between low- and high-severity fire. For
123 low severity fires soil heating is negligible and the impact on soil cover is minimal: therefore,
124 overland flow and soil erosion are reduced compared to high severity wildfires; conversely, in areas
125 affected by high-severity fires, large amounts of fuel are burnt and soil can reach very high
126 temperature (up to 600-800 °C): the impacts on soil hydrology can be extremely negative, such as
127 strong water repellency and very low infiltration capacity [13] (Figure 1). Overall, soil burn severity
128 is considered as a key descriptor of the magnitude of the changes in the soil for its implications on
129 both the hydrological response and vegetation recovery [27].
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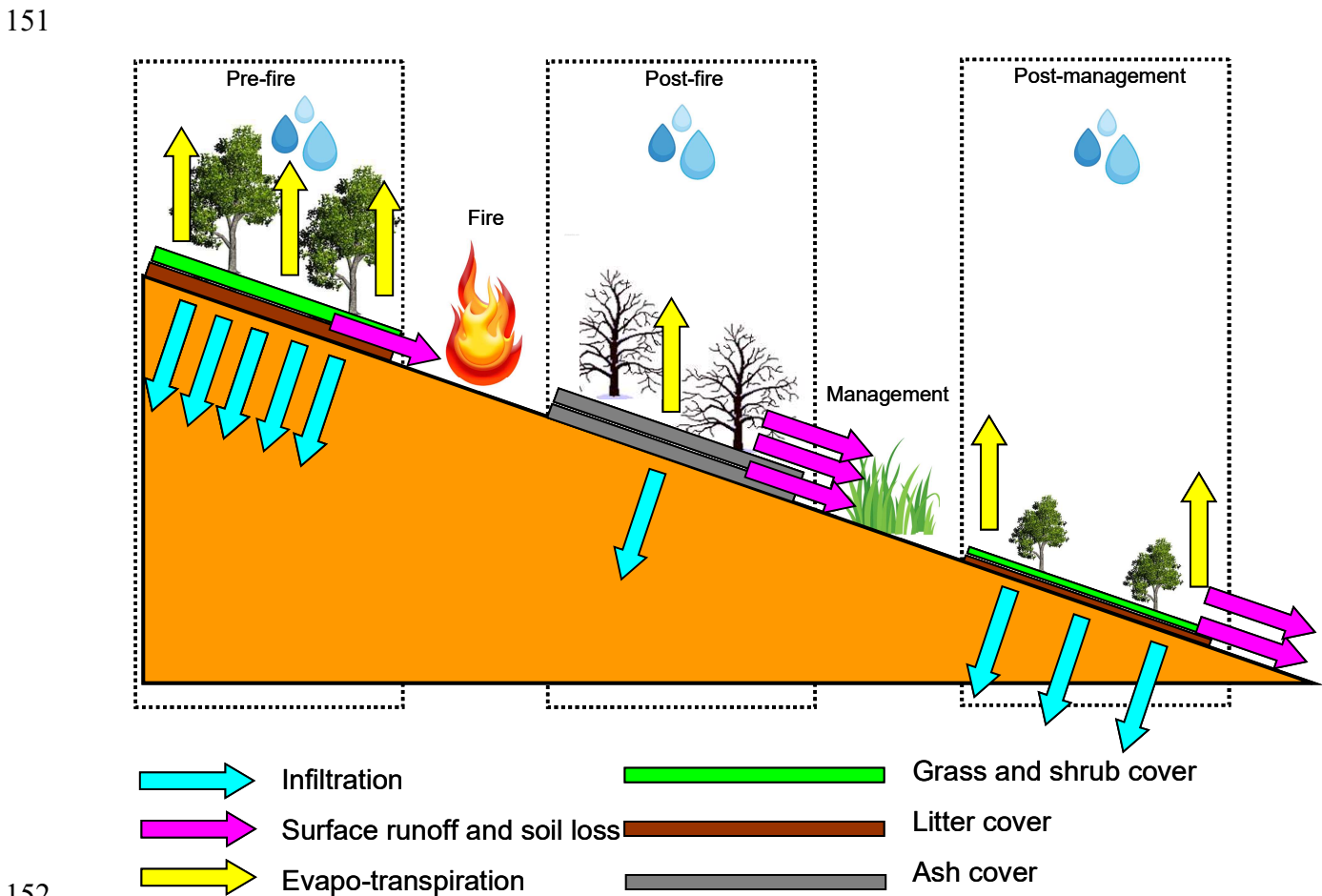
Notes: * Changes in organic matter (OM) content depend on fire severity, since OM can increase in low-severity fires or can decrease after fires with high severity.

135 ** Changes in soil water repellency and aggregate stability depend on fire severity (soil water repellency: no changes
 136 under 200 °C, increases between 250 and 300 °C, and disappearance over 300-400 °C; soil aggregate stability: no
 137 changes or slightly increases under 220 °C, decreases between 380 and 460 °C, and aggregate destruction over 460 °C,
 138 except for clayey soils, which can show increased aggregate stability at high temperatures);
 139 *** green, brown and blue-coloured boxes refer to the three environmental domains affected by wildfire (vegetation,
 140 soil and water, respectively).

141
 142 Figure 1 - Scheme of wildfire effects on soil and vegetation with implications on soil hydrology.
 143

144 **3. Monitoring of post-fire management strategies on soil hydrology**

145
 146 The need to mitigate the fire impacts on soil’s hydrological response has increased the use of post-
 147 fire treatments, whose effects have been largely experimented mainly in the United States, Australia
 148 and Europe [28,29]. The objectives of post-fire management are flood control, reduction in soil loss
 149 and sediment yield, restoration of the ecological functions, and management of the residual fuels to
 150 mitigate the future wildfire risks [14,30] (Figure 2).



152
 153 Figure 2 - Scheme of the hydrological processes acting before and after wildfire with the effects of
 154 post-fire management (adapted from [31]).

155

156 Post-fire treatments can be adopted as emergency or restoration strategies to reduce soil degradation
157 and control floods in fire-affected areas. These treatments can be practised both on hillslopes and
158 river channels. Hillslope treatments (e.g. afforestation, seeding, mulching, salvage logging, erosion
159 barriers - the latter including log erosion barriers or contour felled log debris - or soil preparation)
160 are targeted to quick restore vegetation and soil cover, remove residual fuel to burn, protect soil
161 from raindrop impact, reduce overland flow, trap sediments, and increase water infiltration [13].
162 Channel treatments (such as the construction of rock or concrete check dams) aim at delaying the
163 flood propagation, reducing the sediment transport in watercourses, and retaining eroded sediments
164 [32].

165

166 *3.1. Afforestation and seeding*

167

168 Afforestation is targeted to increase the vegetation cover and improve the soil hydrological
169 properties in areas affected by wildfires. However, the artificial vegetation cover is often less dense
170 and stable compared to natural reforestation. After being seeded, the new plants need time to
171 properly establish in new field conditions, being soil cover by new plants delayed for some years.
172 Although being generally successful, in some cases post-fire afforestation may lead to increased
173 runoff and erosion, because some tree species (such as pines) increase soil water repellency;
174 furthermore, the use of highly flammable trees increases the risk of future frequent wildfires [15].

175

176 *3.2. Mulching*

177

178 In agricultural lands, rangelands, fire-affected areas and anthropic sites, mulching is widely used to
179 limit the negative impacts of surface runoff and soil erosion [29,33]. Mulching is targeted to
180 increase infiltration rates, ground cover and soil quality, if used properly and at the correct time
181 [11]. This treatment consists of dispersing on the soil surface organic and inorganic materials as an
182 alternative surface cover, such as agricultural straw, plant leaves, plastic film, logging slash,
183 shredded barks, wood strands, chips and shreds, as well as gravel and loose soil [34,35]. Among the
184 different mulch materials, vegetal residues are considered the most effective to reduce the soil
185 hydrological response also in agricultural soils [33]. In general, organic residues, such as straw and
186 wood residues, are preferred to other mulch materials, due to its wide availability, high soil
187 covering capacity, low cost and easy-of-handling [4]. Agricultural straw can be applied from the air
188 (heli-mulching), thus allowing the treatment of extensive burned areas in a relatively short time

189 [36]. The selection of the mulch material depends on the local availability and effectiveness, but
190 also on the area to be treated. Large burnt areas with difficult access are more expensive to be
191 treated, forcing the adoption of lower application rates [34,35].

192 Hydrological response of soil is effectively reduced by mulching, thanks to three effects: (i)
193 increase in interception of raindrops and thus reduction in rain splash detachment; ii) reduction in
194 surface sealing and crusting, thereby increasing infiltration; and (iii) increase in surface roughness
195 and slowdown of velocity of overland flow that results in lower soil detachment by sheet and
196 concentrated flows [35,37]. These effects of mulching help stabilizing soil, reducing sediment
197 movement, preventing the loss of soil productivity and preventing the risk of flooding [38].
198 Generally speaking, soil mulching after wildfire is more effective against erosion than its impact on
199 runoff discharge [29].

200 Several factors influence the effectiveness of mulching on soil hydrology (e.g., resistance to
201 physical degradation, strand length, thickness of the application), but the application rate and
202 ground cover are considered as the most important factors [34,35]. Application rates range from 1
203 to 11 Mg ha⁻¹ with straw and forest residue mulches for post-fire unlogged environments until 20-
204 fold values for skidder-compacted soils [35]. More specifically, with regard to the ground cover, a
205 mulch cover rate of 80% using straw and forest residues have been reported to decrease runoff and
206 erosion by 50% and 80-90%, respectively [35,37]. Short-term studies showed that wheat straw
207 mulch treatment reduced the erosion rates by 50-99% in the first two post-fire years at 70% or more
208 ground cover [20]. Concerning to the mulch application rate, straw mulch at rates of 0.6 to 1 Mg ha⁻¹
209 ¹ reduces soil losses by 45-70%, while under forest residue mulch at rates of 1 to 2.6 Mg ha⁻¹ soil
210 erosion decreases by 65-90% during the first post-fire year [4]. As a general rule, it is widely
211 assumed that mulching at rates of 2-3 Mg ha⁻¹ above 60% ground cover can significantly reduce
212 post-fire stream flow and soil loss [20,35].

213 The presence of a mulch cover on soil is able to increase soil moisture, by slowing runoff and
214 increasing infiltration, and reduce topsoil temperature fluctuations and evaporative loss, by
215 impeding solar radiation reaching the soil surface. Such impacts on soil moisture and temperature
216 support the ability of some plant species to germinate and establish [10,33,39,40]. In addition,
217 mulching enhances and fastens tree regeneration without influencing plant diversity
218 [41]. Therefore, mulch application has the potential to change many aspects of the post-fire
219 environment and therefore mulching has become one of the most direct and effective soil restoring
220 techniques after wildfire [29,39].

221 However, mulching can also have negative effects as post-fire treatment. In some cases, compared
222 to untreated soils, straw may reduce the soil hydraulic conductivity under unsaturated conditions,

223 particularly in the drier season: this suggests caution in mulch use in the case of heavy storm
224 occurrence in summer [42]. Straw mulch can be displaced by wind, which can leave some slopes
225 bare and too much thick layer of straw in other areas, the latter preventing the emergence of
226 vegetation due to the sunlight absence [20]. Moreover, agricultural straw may contain seeds,
227 chemicals and parasites, which can be the sources of non-native vegetation and plant diseases. A
228 possible solution may be the use of mulches from forest materials (e.g. wood strands, chips or
229 shreds), which are less likely to carry non-native seeds chemical residues, and show greater
230 resistance to wind displacement. The forest materials provide similar protection from erosion at
231 equal ground cover rates and show a longer effectiveness, although requiring higher application
232 rates compared to agricultural straw [38]).

233 Moreover, the research experiences that are proven the effectiveness of mulching as post-fire
234 management treatment have been carried out mainly at the plot scale; therefore, the extent at which
235 the application rates can be effective and the quantification of hydrological benefits on larger scales
236 are still unknown [34]. This means that further research is needed to quantify these key issues for a
237 consolidated use of mulching, especially in the areas where soil erosion by water represents a severe
238 threat, as the areas affected by wildfires [33].

239

240 *3.3. Salvage logging*

241

242 Salvage logging as post-fire management technique is based on the removal of dead and damaged
243 trees, often carried out using machinery that drags the wood over the burned soil. This treatment is
244 generally executed in the first two years after wildfire, in order to recovery the economic value of
245 the wood and to reduce the danger of another wildfire [20]. Salvage logging can sometimes be
246 prolonged, when the recovered wood must be used for other purposes, such as wildfire wood or log
247 home [2].

248 Generally speaking, salvage logging heavily impacts on soil properties, particularly in the short-
249 term after torrential rainfalls, determining soil degradation [43]. The hydrological effects of salvage
250 logging can be contrasting. This technique may be beneficial, if the water repellent layer is broken
251 up - and thus the infiltration is increased - as well as the addition of logging slash is able to increase
252 the surface cover, with consequent decreased erosion rates. Conversely, salvage logging can have
253 heavy hydrological impacts, since the machinery and logging equipment exert a high pressure on
254 soil, with consequent compaction and rut formation from logging traffic [11,20]. Skid trails formed
255 by machine wheels create preferential flow paths, where surface runoff may concentrate and its
256 sediment transport capacity increases [44]. Post-wildfire salvage logging is able to increase flooding

257 and soil erosion over an area by two orders of magnitude compared to the undisturbed soil [20].
258 Moreover, the disturbance of salvage logging can affect soil properties and vegetation regeneration
259 for decades [15]. A possible countermeasure may be mitigating the increased erosion rates with
260 post-fire treatments (e.g., distributing available treetops and branches on skid trails and landings), to
261 protect the logged areas from surface erosion [20]. The impacts of salvage logging on vegetation are
262 also uncertain, as shown by the very low vegetative re-growth rates recorded immediately after a
263 wildfire compared to the long-term regeneration [44].
264

265 *3.4. Erosion barriers*

266

267 Erosion barriers have been used for decades to stabilize hillslopes and mitigate post-wildfire runoff
268 and erosion. These barriers are made of inert or vegetal materials, the latter having the advantage of
269 being biodegradable. The erosion barriers are classified into contour-felled logs, straw wattles,
270 contour trenches, straw bales, fascines, vegetal strips and buffers [16]. The use of erosion barriers
271 prevents sediment delivery to downstream water bodies by slowing down runoff, causing localized
272 ponding, and trapping sediments [15]. Some erosion barriers have been shown to have a sediment
273 trapping rate of 40% or more, and resulted to be very cheap compared to other hillslope
274 stabilization techniques [16]. Contour-felled logs were found to be even very effective to reduce
275 runoff and sediment yield on burned forest subject to machinery salvage logging, provided that the
276 barrier distance is higher than 20 m [40]. Literature reports also some cases in which erosion
277 barriers have not been successful in reducing post-fire soil erosion, presumably due to the excessive
278 burn severity of the wildfire or defective construction (e.g., [45]). The decrease in post-fire erosion
279 rate over time is also attributed to the soil improvement and vegetation recovery [45]. It has been
280 demonstrated the beneficial effects of log erosion barriers and contour-felled log debris on promote
281 soil multifunctionality and plant diversity to recover community-level properties and forest
282 functions also in the short term after wildfire [46]. In this sense, log erosion barriers are slightly
283 more effective in improving soil quality and vegetation regeneration compared to contour-felled log
284 debris, and these beneficial effects help retaining sediments and limiting nutrient loss, which is
285 essential to recover vegetation after a wildfire [28].

286 However, this technique is mainly effective for low-intensity rainfalls, since the barriers easily get
287 overtopped after high-intensity rains; furthermore, they lose their effectiveness due to the
288 progressive sediment accumulation and material degradation, if not maintained or regularly cleaned
289 [15,40]. Research still has to quantify the direct impacts of the different types of erosion barriers,
290 since this technique has been generally adopted in combination with other post-fire management
291 strategies (grass and vegetative barriers or contour-felled logs) [16].

292

293 *3.5. Soil preparation*

294

295 Soil preparation (e.g., by tillage, conditioning and terracing) is considered as a viable practice to
296 reduce the hydrological response of soil, particularly in croplands (e.g., [47–49]). However, when
297 used for post-fire management, these techniques may be less beneficial for improving soil quality
298 and hydrology (e.g., [37]) and furthermore expensive, when very large forest areas must be treated.
299 As a matter of fact, soil tillage is theoretically able to break up fire-induced soil water repellency or
300 sealing and increase infiltration [20,38]); at the same time, tillage can decrease organic matter
301 content and worsen structure degradation of soil, leading to a decreased infiltration (at least in the
302 short-time) and resulting ineffective to reduce runoff and erosion [50]. Soil tillage may synergistic
303 with other post-fire techniques, since this can increase the capacity of limiting runoff and erosion in
304 same cases by 20% [37]; in this direction, future research must watch at the integration of the most
305 effective post-fire management techniques with soil pre-treatments that are able to increase the
306 hydrological restoration capacity of burned areas.

307

308 *3.6. Other techniques*

309

310 The success of natural fibre webs for supporting vegetation growth and soil stabilization in
311 degraded hillslopes, highway and railway embankments and construction sites has recently
312 suggested the use as mats and rolls made of coconut fibres as post-fire restoration techniques.
313 Coconut fibre webs have been found to delay the time to runoff generation, enhance soil infiltration
314 capacity, decrease splash erosion and reduce the velocity of overland flow [15]. The application of
315 moss crust, which is a very fast soil colonizer after wildfires, may also be beneficial in the first
316 periods after wildfire. Although the runoff response of the soil increases, sediment and organic
317 matter losses can be reduced by over 60% and 30%, respectively, especially during the rainiest
318 season [51]. Moreover, mosses are important for restoring soil functionality after high-severity fires
319 (particularly in terms of fertility and microbial activity), thus improving post-fire vegetation
320 recovery [52].

321

322 *3.7. Check dams*

323

324 The use of check dams is presumably the most common measure to control soil erosion in channels,
325 particularly under the semi-arid conditions [32,53]. This technique has been proposed and applied

326 since long time to trap sediment and stabilize channels also in burned catchment of different size
327 and conditions. Check dams are built of rock, wood, straw bales, rock gabions or a combination of
328 these materials in ephemeral channels draining low-order catchments [54]. In general, the sediment
329 storage capacity of check dams directly depends on its height and channel slope [55,56]. As post-
330 fire management strategy, the use of straw bale check dams is largely diffused in USA, due to their
331 quick and easy installation in burned catchments. However, the effectiveness of this channel
332 treatment is questionable for three main reasons. First, the treatment is unsuccessful in primary
333 watersheds or small catchments, since fine sediments and ashes wash permeate the structures and
334 are released into higher-order channels. Second, the straw bale check dams are prone to an easy
335 failure, because of piping and stream dragging. Third, once the check dams are filled with
336 sediments, their storage capacity is depleted, and this occurs just after two years post-fire. Overall,
337 the installation of light check dams such as straw bales is viable only in areas with low rainfall
338 intensities and soil with low erodibility [54]. However, research about the effectiveness of check
339 dams in burned catchments has not been sufficient to explore the large variability of climatic,
340 geomorphological and vegetation conditions, in order to deeply assess the contribution of channel
341 structures as catchment level. It is likely that check dams are able to reduce the catchment
342 connectivity, but their effectiveness in reducing the runoff and erosion rates is less pronounced
343 compared to hillslope treatments, once the water and sediment flows are already mobilised across
344 the torrent system.

345

346 **4. Modelling post-fire management impacts**

347

348 In areas affected by high-severity fires, accurate hydrological predictions using computer-based
349 models help land managers in the adoption of the most suitable actions to mitigate post-fire land
350 degradation and rehabilitation planning [57,58].

351 An ample and eminent literature exists about modelling experiences in burned forests [59]. To
352 summarise, simple empirical and semi-empirical models (such as the Universal Soil Loss Equation,
353 USLE, the Morgan–Morgan–Finney model, MMF and the revised versions [31,57,60], and
354 physically-based models (such as the Water Erosion Prediction Project, WEPP, the Pan-European
355 Soil Erosion Risk Assessment, PESERA, the Soil and Water Assessment Tool model, SWAT) have
356 been widely tested with results generally showing sufficient reliability and accuracy [58,61]. These
357 modelling experiences have relied on adaptations of existing hydrological models to fire-induced
358 changes [31,62]. A recent review of Lopes et al. [63] reports that 73% of the related case studies
359 involved model adaptation to burned conditions. The existing models do not usually account for the

360 impacts of wildfires on vegetation cover and soil properties [64]. In general, fire-adapted
361 algorithms, methods to parameterize post-fire vegetation and soil properties, empirical “fire factor”
362 or adjustments of input parameters - such as ground cover, surface roughness or soil hydraulic
363 properties - have been proposed for the existing models (e.g., [31,58,60,63]). According to the latter
364 authors, only 21% of the reported studies attempted to accommodate new processes and 27% of the
365 papers have tested the accuracy of the models in simulating the effects of mitigation measures.
366 Referring only to the most recent experiences, only Nunes et al. [64], using SWAT to simulate post-
367 fire afforestation, Pastor et al. [65], working with the long-term soil erosion “LandSoil” model
368 under post-fire mulching, combined with riparian vegetation maintenance/restoration and reduced
369 tillage, as well as Vieira et al. [31], comparing MMF, RUSLE and PESERA models for mulched
370 areas, and Zema et al. [57], using an adapted version of MMF in post-fire pine stands treated with
371 mulching, have evaluated the accuracy of these hydrological models to simulate soil hydrology
372 under post-fire management measures. Vieira et al. [31] concluded that the RUSLE model seems to
373 be ideal for prioritization of areas prone to wildfire risks, mainly due to its simplicity and reduced
374 data requirements, while the more complex MMF and PESERA models are more suitable for
375 testing different land management scenarios. Zema et al. [58] have adopted a novel approach to
376 post-fire hydrological modelling, proposing an Artificial Neural Network to predict with very
377 satisfactory results surface runoff and soil erosion after wildfire under Mediterranean climate
378 conditions.

379 Overall, it can be concluded that the available literature about post-fire hydrological modelling is
380 not homogeneously distributed worldwide, in accordance with Lopes et al. [63]. Erosion modelling
381 is well developed in the U.S.A., where post-fire prediction models are commonly applied, but in
382 other regions, where the hydrological processes may be site-specific, research is still far from being
383 exhaustive [60]. For instance, Mediterranean burnt areas have intrinsic conditions (e.g., very
384 shallow soils, strong soil water repellency, peculiar hydrologic regime). In these environments, the
385 available hydrological models, developed in other climatic contexts and not in fire-affected areas,
386 may find limited applicability and therefore require targeted modifications [31]. As previously
387 outlined, the hydrology of burned areas is extremely complex, due to the large number of
388 influencing factors, and the post-fire management techniques are numerous and different in nature.
389 Therefore, the simulation of the hydrological response of burnt and treated soils is a challenging
390 task, which requires further research [61]. The statement by Lopes et al. [63] can be shared, given the
391 evidences that such studies should alternatively adapt the existing models to the hydrology of
392 burned and treated soils or develop new prediction tools under variable post-fire conditions and
393 management [63]. The exploitation of powerful analytical techniques, such as the remote sensing, to

394 derive ground cover and soil properties (e.g., water content and burn severity maps) from soil
395 survey maps may enhance the prediction accuracy and the easiness of use of these tools.

396

397 **5. Future perspectives and conclusions**

398

399 The intense research and discussions about the hydrological effects of post-fire management is still
400 open, and even some uncertainty remains whether some areas should recover naturally or it is more
401 convenient to implement soil and vegetation restoring measures [14,30]. On this context, the
402 discussion above has demonstrated the general effectiveness of hillslope and channel measures after
403 wildfire on both soil and vegetation components.

404 However, despite the large availability of effective post-fire management techniques, it is clear that
405 the optimal solution in fire-affected areas remains a combination of these actions. In other words,
406 the post-fire management techniques should be synergistically integrated in a holistic approach that
407 has to consider the integrated hydrological response of the watershed system [66]. Moreover,
408 because of the wide-ranging temporal and spatial effects of each management technique, upscaling
409 the research from the hillslope or channel scale to the watershed scale is warmly recommended, in
410 order to better understand the effects of each technique at large spatial scales [11]; their effects after
411 years or decades must not be neglected rather than limiting the analysis to the short-time impacts.

412 Despite the large body of studies that have evaluated the effectiveness of the post-fire management
413 techniques in a wide range of climatic, geomorphological and ecological conditions, more
414 experiences are needed to identify the most effective strategy, which should be tailored to site
415 and wildfire characteristics [15]. In this sense, comparative studies of more than one technique
416 against the negative hydrological impacts of post-fire management are welcome. These studies
417 would give as support the scientific evidence about the effectiveness of each action in a territory of
418 specific characteristics. However, there is also a need to develop methods of feasibility assessment
419 using suitable indicators about the sensitivity of each environmental context to wildfire and the
420 benefits of the different techniques. This assessment can help prioritising the management actions in
421 restoration projects, due to the fact that the cost of mitigation measures may be prohibitive over
422 very large fire-affected areas [67].

423 Regarding the prediction of the hydrological effects of post-fire management using computer
424 models, future research paths should go towards a large applicability of post-fire models, which
425 currently is not homogeneously distributed worldwide. The existing models should be adapted as
426 much as possible to burned conditions with attention to the impacts of soil changes in runoff and
427 erosion generation mechanisms. Many studies have limited their evaluations to existing models

428 under burned and unburned conditions; research should go ahead, focusing the effects of post-fire
429 management actions, since these types of modelling experiences are generally scarce and confined
430 to very few environments and techniques. Finally, future modelling studies should include
431 uncertainty analysis, in order to give modellers the level of reliability and accuracy of their
432 hydrological predictions [63].

433

434 **References**

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