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17 **Short-term changes in soil functionality after wildfire and straw mulching in a *Pinus***  
18 ***halepensis* M. forest**

19

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31

32 **Abstract**

33

34 Understanding the changes in physico-chemical and microbiological soil properties induced by  
35 wildfire and post-fire soil restoration technique (e.g., soil mulching with straw) is very important  
36 in the Mediterranean environment, where the forest ecosystems are particularly prone to erosion  
37 and degradation risks. Nevertheless, the studies about the effects of straw application on  
38 functionality of burned soils in Mediterranean forest ecosystems are scarce. To fill this gap, this  
39 study has evaluated the seasonal changes (from spring to autumn) in important physical, and  
40 chemical soil properties and enzymatic activities in burned (treated with mulching or not) plots,  
41 compared to non-burned soils, after a wildfire occurred in a *Pinus halepensis* M. forest. The

42 monitoring activity has confirmed that the treatment of burned soils with straw mulching  
43 improves its functionality in the short-term, assumed as working hypothesis. More specifically,  
44 compared to non-burned soils, although soil pH was stable and the electric conductivity  
45 noticeably reduced the organic matter content increased and the soil C/N ratio recovers in one  
46 year in burned and mulched soils. The increases of basal respiration as well as microbial carbon  
47 and glomaline contents after mulching indicated higher activity of soil microorganisms and  
48 increased carbon and nitrogen storage. Moreover, all the microbiological and enzymatic  
49 activities improved, except for dehydrogenase activity. Finally, the Canonical Analysis of  
50 Principal Coordinates confirmed the differentiated functionality of non-burned, burned and non-  
51 treated, and burned and straw-mulched soils. Overall, this study highlights that soil functionality  
52 of wildfire-affected areas significantly benefits with straw mulching treatment, which could be  
53 adopted as countermeasure against soil quality decay in the Mediterranean forest ecosystems.

54

55 **Keywords:** High-severity fire; Mediterranean forest; soil enzymes; soil respiration; soil organic  
56 matter.

57

## 58 **1. Introduction**

59

60 Wildfires are a natural disturbance factor in Mediterranean forest ecosystems, where climate  
61 change and fire suppression have altered natural fire patterns (Kaufman et al., 2001). High-  
62 intensity fires modify the hydrologic response of soil and enhance its degradation, removing  
63 vegetation and altering chemical, physical and biological soil properties (DeBano, 2000). For  
64 instance, as regards soil hydrology, it is well documented that decreased infiltration and

65 increased overland flow after wildfires, leading to increasing erosion rates and soil degradation  
66 (e.g., Robichaud and Waldrup, 1994). Therefore, mitigation of post-fire effects is compulsory in  
67 order to reduce the soil exposure to hydrological and quality degradation. Mitigating the post-fire  
68 effects on soil has resulted in the increased use of post-fire treatments, in which soil stabilization  
69 treatments are crucial for diminishing soil degradation (Gómez et al., 2019).

70 Post-fire treatments may be divided into three categories: (i) emergency stabilization; (ii)  
71 rehabilitation; and (iii) restoration (Lucas-Borja et al., 2019). Many experiments developed in  
72 USA and Europe have shown that long-term rehabilitation and restoration actions are often  
73 focused on the biotic components of the ecosystem (Hessburg and Agee 2003; Beschta et al.  
74 2004; Robichaud, 2005; 2010; Fernandez and Vega, 2016; Gómez et al., 2019; Lucas-Borja et al.,  
75 2019). For these activities, recovery of native plant communities and habitats, maintenance of  
76 plant biodiversity, re-establishment of timber or grazing species and control of invasive weeds  
77 are the most important targets. As regards the emergency stabilization actions, mulching is  
78 considered as one of the most efficient treatment to stabilize the soil of the burned area and to  
79 reduce additional damage to soil and vegetation immediately after wildfire. This treatment  
80 consists in spreading organic material (e.g., wheat straw or woodchips) over soil immediately  
81 after a fire and just before the first autumn rainfall. The benefits of mulching have been largely  
82 demonstrated in literature (e.g. Prosdocimi et al., 2016). Strictly speaking about the hydrological  
83 aspects, Smets et al. (2008) have shown that mulching provides a suitable soil cover that reduces  
84 raindrop impact, prevents soil sealing, promotes infiltration and slows runoff. Therefore, post-  
85 fire mulching is critical for reducing runoff and soil erosion, especially after clear cutting in areas  
86 affected by crown-fire, where the soil is exposed to the rainfall action and the amounts of  
87 logging debris on the soil surface may be low (Lucas-Borja et al., 2019).

88 Beside these benefits, some problems using straw or woodships mulching as post-fire emergency  
89 treatment, such as, for example, straw blowing under strong winds, quick decomposition or  
90 emergence of non-native plant species (Cerdà et al., 2016; Prosdocimi et al., 2016). Luna et al.  
91 (2018) found that, despite woodships mulch was appropriate for reducing erosion and runoff in  
92 restored soils, this type of mulch did not favour vertical water movement towards deeper  
93 horizons and then was not useful in order to increase soil water storage. In general, mulching  
94 may alter soil moisture and temperature, since the mulch layer can obstruct emerging natural and  
95 seeded vegetation by sunlight interception or plant recovery (Lombao et al 2014). Moreover,  
96 straw mulching can generate changes in soil properties, since straw can act as a new source of  
97 vegetal material to be incorporated into the soil. On this regard, organic matter, microbial  
98 biomass carbon, respiration, enzymatic activities or nutrient content of soil, directly or indirectly  
99 linked to vegetal input into the soil (Doran and Parkin 1994; Larson and Pierce 1994; Entry and  
100 Emmingham, 1998; Bastida et al., 2007; Hedo et al., 2015), can be influenced by straw doses  
101 applied to soil with mulching.

102 In spite of this close linkage between physico-chemical and microbiological properties of soil  
103 and mulching, little is known about the effects of straw application on soil functionality of  
104 Mediterranean forest ecosystems, although its positive influence on soil hydrology is well  
105 documented. As far as now, several enzyme activities, specifically related to the cycles of N, P,  
106 C and S (urease, alkaline and acid phosphatase,  $\beta$ -glucosidase and arylsulfatase, respectively)  
107 and some general microbial indicators, such as dehydrogenase activity and soil respiration, have  
108 been proposed as specific indicators of soil functionality (Bastida et al., 2008; Lucas-Borja et al.,  
109 2011; Hedo et al., 2015). In addition, the C/N ratio (Lucas-Borja et al., 2012; Hedo et al., 2015),  
110 soil pH (Lucas-Borja et al., 2012), soil texture (Fterich et al., 2014), nutrients status (Burgess and

111 Wetzels, 2000; Santa-Regina and Tarazona, 2001) or microbiological communities (Wu et al.,  
112 2013) have been used as meaningful indicators of soil functionality. In spite of this knowledge,  
113 more research is needed to better understand whether and to what extent soil functionality is  
114 influenced by straw mulching, with particular reference to the Mediterranean forests, where soils  
115 are particularly prone to erosion and degradation and the fire risk is very intense. Plant and soil  
116 cover may affect the equilibrium of these ecosystems, which in consequence could alter soil  
117 properties and functionality. In these delicate ecosystems, soil functionality plays an important  
118 role in soil fertility and stability by enhancing growth and proliferation of microorganisms,  
119 which accomplish reactions to release soil nutrients for vegetation development (Hannam et al.,  
120 2006). Forest managers and policy makers should know more deeply how straw mulching may  
121 affect soil functionality in wildfire-affected areas to establish proper management guidelines  
122 (Gómez et al., 2019).

123 This study aims to determine whether post-fire straw mulching alter specific indicators of soil  
124 functionality in the short-terms after a wildfire in a Mediterranean forest of *Pinus halepensis* M.  
125 More specifically, straw was applied as mulching treatment immediately after the wildfire in  
126 different areas and then soil microbiological properties were monitored throughout one year in  
127 spring and autumn. We hypothesized that the straw mulching may enhance soil functionality in  
128 the short-term, because it increases the soil organic matter content, which plays an important role  
129 in controlling its metabolic processes.

130

## 131 **2. Methods**

132

### 133 *2.1. Study site*

134

135 The study was carried out in the Sierra de las Quebradas forest (Liétor, Castilla-La Mancha  
136 region, province of Albacete, Central Spain (W1°56'35.02''; N38°30'40.79)) (Figure 1).  
137 Elevation ranges between 520 and 770 m and the aspect is W-SW. The climate of the area,  
138 located on the meso-mediterranean bioclimatic belt (Rivas-Martínez et al., 2002), is semi-arid,  
139 "BSk" according to the Koppen classification (Kottek et al., 2006). The mean annual temperature  
140 and precipitation are 16.6°C and 321 mm, respectively. According to the historical data (1990-  
141 2014) provided by the Spanish Meteorological Agency (AEMET), the maximum precipitation is  
142 concentrated in October (44.5 mm) and the minimum in May (39.6 mm); from June to  
143 September a hot and dry period (air relative humidity below 50%) occurs. According to the Soil  
144 Taxonomy System, soils are *Calcid Aridisol*, with a sandy loam soil texture. Vegetation belongs  
145 to the *Quercus cocciferae-Pinus halepensis* S. series, with a tree cover of Aleppo pine and a shrub  
146 layer of kermes oak (Peinado et al., 2008). The current vegetation of the forest area mainly  
147 consists of *Pinus halepensis* M. stands. In the study site the mean density and height of forest  
148 trees before the wildfire were about 500–650 trees/ha and 7–14 m, respectively. The main shrubs  
149 and herbaceous species were *Rosmarinus officinalis* L., *Brachypodium retusum* (Pers.) Beauv.,  
150 *Cistus clusii* Dunal, *Lavandula latifolia* Medik., *Thymus vulgaris* L., *Helichrysum stoechas* L.,  
151 *Stipa tenacissima* (L.), *Quercus coccifera* L. and *Plantago albicans* L. The use of such species  
152 was an economic driver of the area from the 17<sup>th</sup> century until the middle of the 20<sup>th</sup> century. Its  
153 progressive abandonment and the reforestation by the local public authorities have shaped a  
154 forest landscape composed of Aleppo pines of natural origin growing in shaded areas and  
155 watercourses.

156

157 2.2. Experimental design

158

159 This study was carried out during 2017 inside a drainage basin of the approximately 700 ha  
160 affected by a wildfire in July 2016. Immediately after the wildfire, one site of about one km<sup>2</sup>,  
161 totally covered by *Pinus halepensis* M. and affected by crown fire (tree mortality of 100%), was  
162 selected for study (Figure 1). In the burned area nine rectangular experimental plots (each one of  
163 20 x 10 m) were randomly installed with their longest dimension along the maximum slope.  
164 Plots were distributed selecting certain sites characteristics, slopes and aspects to ensure  
165 comparability among the nine plots used in this study. Distance between plots was always higher  
166 than 200 m. Soil burn severity, measured using the methodology proposed by Vega et al (2013)  
167 and Fernandez et al (2017), was high in each plot, thus allowing to compare our experimental  
168 plots. A weather station (WatchDog 2000 Series model), purposely placed in the study area  
169 during the study period, measured precipitation depth and intensity, and air temperature (Table  
170 1). Three of the nine experimental plots were placed in an unburned area, one km away from the  
171 burned site, and assumed as control. Three other plots were located in the burned area, but not  
172 treated.

173 Mulching treatment was assigned in September 2016 to the remaining three plots located in the  
174 burned area. Mulching consisted of manual application of straw (0.2 kg/m<sup>2</sup> of dry weight) on  
175 plot soils at an initial depth of three centimetres. This dose was proposed by different authors to  
176 achieve a cover over 80% in plots located in the north of Spain (Vega et al., 2014). Moreover,  
177 such amount of straw is also successfully (the biophysical point of view) used in agricultural  
178 land affected by intolerable erosion rates (Cerdà et al., 2017). To summarize, three replicated  
179 plots were non-burned soils (and hereinafter indicated as "NB"), three plots were burned and



180 non-mulched soils (three replicates, hereinafter "B+NM"), and three plots were burned and then  
181 mulched soils (three replicates, "B+M"). Prior to soil sampling, the percentages of vegetation  
182 cover, rock fragments, dead matter, bare soil and ash on the plots were measured one day after  
183 the mulching application (in September 2016), in the mid of the study period (in March 2017)  
184 and at the end of the experiment (in July 2017). More details about soil cover measuring methods  
185 and results are reported in [Lucas-Borja et al. \(2019\)](#).

186

### 187 *2.3. Soil sampling and analyses*

188

189 As regards soil sampling, three soil samples (each of 600 g) were collected in each plot in two  
190 seasons (May 2017 and November 2017) throughout one year after the wildfire for a total of 18  
191 soil samples, 3 treatments (NB, B+M, B+NM) x two seasons (autumn and spring) x three  
192 replicates. Soil samples consisted of the composition of further six sub-samples, randomly  
193 distributed over each plot, in order to take into account the spatial variability of plot soils. Each  
194 soil sample was collected from the upper soil layer (depth of 5 cm) after litter removal, then  
195 sieved (at 2 mm) and kept at 4 °C. Soil analyses were carried out 1 day after sampling. Sampled  
196 soil was analysed for the main physical, chemical and microbiological properties. Concerning the  
197 physical and chemical properties, texture (soil contents of sand, silt and clay) was analysed  
198 according to the methods by the method of [Gutián and Carballás \(1976\)](#). Soil pH and electrical  
199 conductivity (EC,  $\mu\text{S}/\text{cm}$ ) were determined in a 1:5 (w/v) aqueous solution by portable analyser  
200 with dedicated probes. Organic matter content (OM, %) was measured by potassium dichromate  
201 oxidation method ([Nelson and Sommers, 1996](#)). Organic carbon (OC, %) was calculated by  
202 dividing OM by 1.72 ([Lucas-Borja et al., 2018](#)). The C/N ratio (-) was calculated according to

203 Lucas-Borja et al. (2012). Total nitrogen (TN, %) was determined using the Kjeldahl (Bremner  
204 and Mulvaney, 1982). As regards the microbiological properties of soils, microbial carbon (MC,  
205 expressed as mg C kg<sup>-1</sup> dry soil) was measured by the fumigation-extraction methods (Vance et  
206 al., 1987). Basal soil respiration (BSR, expressed as the CO<sub>2</sub> rate (µg hour<sup>-1</sup> g<sup>-1</sup> of dry soil) was  
207 determined in a multiple sensor respirometer (Micro-Oxymax, Columbus, OH, USA). Soil  
208 dehydrogenase activity (DHA, expressed as µg INTF hour<sup>-1</sup> g<sup>-1</sup> of dry soil) was determined as  
209 the reduction of p-iodonitrotetrazolium chloride (INT) to piodonitrotetrazolium formazan using  
210 the modified method of Von Mersi and Schinner (1991). Urease activity (UA, expressed as µmol  
211 N-NH<sub>4</sub><sup>+</sup> hour<sup>-1</sup> g<sup>-1</sup> of dry soil) was measured according to the method of Tabatabai, (1994),  
212 using urea as substrate and borate buffer (at pH = 10) (Kandeler and Gerber, 1988). Acid  
213 phosphatase (Acid-PA) and β-glucosidase (BGA) activities - both expressed as µmol pNP hour<sup>-1</sup>  
214 g<sup>-1</sup> of dry soil - were determined according to the methods of Tabatabai and Bremner (1969) and  
215 Eivazi and Tabatabai (1977), respectively. Glomalin-Related Soil Protein (GPRS, expressed as  
216 g<sup>-1</sup> dry soil) content was evaluated according to Lozano et al. (2016).

217

#### 218 2.4. Statistical analyses

219

220 Statistical differences on physical, chemical and microbiological soil variables of non-burned  
221 spring (NB-spring), non-burned autumn (NB-autumn), burned and non-mulched spring (B+NM-  
222 spring), burned and non-mulched autumn (B+NM-autumn), burned and mulched spring (B+M-  
223 spring) and burned and mulched autumn (B+M-autumn) samples were evaluated with univariate  
224 and multivariate Permutational Analysis of Variance (PERANOVA and PERMANOVA,  
225 Anderson, 2001) using a three-factor design: (i) fire occurrence, (ii) mulch addition, (iii) season

226 of the year. To study the relationships between these soil properties was used a Pearson's  
227 correlation analysis and to assess the similarities among the soils samples of each treatment was  
228 used a Canonical Analysis of Principal Coordinates (CAP) after normalizing the data. The CAP  
229 analysis is a constrained nonparametric ordination procedure, widely used as ecology ordering  
230 method, since it allows the use of any distance or dissimilarity measure, and, at the same time,  
231 takes into account correlation structure among response variables (Anderson and Willis, 2003).  
232 This analysis consists of the following steps: (i) Principal Coordinate Analysis (PCA) on the  
233 data matrix  $\mathbf{Y}$ , using a similarity measure (in this study using Euclidean distance), which yields  
234 orthogonal  $\mathbf{Q}$ ; (ii) selection - based on in minimum misclassification error or minimum residual  
235 sum of squares - of an appropriate number of axes  $m$  as a subset of  $\mathbf{Q}$ , thus defining a matrix  $\mathbf{Q}_m$ ;  
236 (iii) application of a traditional canonical analysis (e.g., a Canonical Correlation Analysis, since  
237 it  $\mathbf{X}$  contains quantitative variables) on the first  $m$  axes of  $\mathbf{Q}$ . The software used for the statistical  
238 analyses was PRIMER V 7® with PERMANOVA add-on (Anderson et al., 2008) and  
239 Statgraphics Centurion XVI® (StatPoint Technologies, Inc.).

240

### 241 3. Results

242

#### 243 3.1. Effects of wildfire and mulching on physico-chemical and microbiological soil properties

244

245 The PERMANOVA analysis showed significant differences ( $p < 0.001$ ) among soils sampled in  
246 NB (spring and autumn), B+NM (spring and autumn) and B+M (spring and autumn) plots (Table  
247 2). The results of CAP evidenced that the soil samples were constricted in the six treatments  
248 analyzed: (i) NB soils sampled in spring field campaign; (ii) B+NM soils sampled in spring; (iii)

**Comento [YYY1]:** No es claro; por supuesto, los mostreos tienen que estar en diferentes clusters.

249 B+M soils sampled in spring; (iv) NB soils sampled in autumn; (v) B+NM soils sampled in  
250 autumn; (vi) B+M soils sampled in autumn. Selecting the first 10 axis of the PCA (that is,  
251 choosing  $m = 10$ ), 99.97% of the variance of the samples was explained and 100% of correct  
252 assignations (12 on 12) of the soil samples in the each cluster of treatments was achieved (Table  
253 3). The results of the cross validation correctly allocated all the observations to original groups  
254 for the choice of  $m$  equal to 10 (Table 4). All clusters were significantly different *each other* with  
255 the exception of the soils sampled in B+NM both in autumn and spring and those sampled in  
256 B+M in autumn. Moreover, the microbiological parameters (BSR, MC and the enzymatic  
257 activities) as well as GPRS and the contents of silt, OM, TN and C/N are mainly oriented to the  
258 clusters grouping the soil samples treated with straw mulching after fire (Figure 2). On the  
259 contrary, DHA, content of clay and pH were oriented to the clusters consisting of *non-burned*  
260 samples. Finally, it is noteworthy that OM and TN content of soils have higher loadings on axe 1  
261 (CAP1), while important microbiological parameters (BGA and UA) have higher weights on axe  
262 2 (CAP2).

Comento [YYY2]: Porqué  
12 y no 18 (3 replicaciones x 3  
treatments x 2 estaciones)?

### 264 3.2. Differences among treatments and temporal changes in physical and chemical soil 265 properties

266  
267 The texture of NB plots was loam-clay-sandy, while both the burned soils (B+M and B+NM)  
268 were sandy-loam both in spring and autumn 2017 (Table 5).

269 While the textural properties of NB soils remained practically constant in time, some significant  
270 changes in textural contents were monitored from autumn to spring in the other experimental  
271 plots. Moreover, compared to the first field campaign in spring, the clay content *significantly*

272 decreased (by 48%) and the percentage of silt simultaneously increased (by 27%) in B+NM  
273 plots. In B+M soils, the percentage of silt significantly increased (by 26%) and the sand content  
274 decreased (by 12%) from spring to autumn (Table 5).

275 In general, most of the physico-chemical properties (contents in OM, OC, TN and EC) were  
276 significantly different among the three analyzed treatments in both field campaigns. As regards  
277 their time evolution, the OM and OC contents were almost stable in NB soils, while they  
278 significantly increased (by 30% both) in B+NM plots and decreased (by 16% and not  
279 significantly) in B+M soils from spring to autumn. In this period, TN increased in NB (by 28%)  
280 and B+NM plots (by 14%) and decreased (by 18%) in B+M soils, although not significantly  
281 (Table 5).

282 Both in autumn and spring, the highest EC, OC and OM contents were detected in B+M plots,  
283 while the lowest values of these properties were found in NB soils, except for EC, which showed  
284 the lowest value ( $81.35 \pm 19.85 \mu\text{S}/\text{cm}$ ) in B+NM soils sampled in autumn. High reductions in EC  
285 values from spring to autumn were found in NB soils (-18%) and mainly in B+M (-57%) and  
286 B+NM plots (in the latter the value practically halved from autumn to spring). EC of NB soils  
287 was significantly different from the values measured in burned (B+NM and B+M) plots in  
288 spring, while in autumn the value recorded in B+M soils become significantly different from the  
289 other treatments. On the contrary, there were no significant differences in soil pH in every season  
290 and treatment. All soils showed always a slightly alkaline pH (on the average in the range 8.4-  
291 8.7) and low variability was found for soil pH between the monitored seasons (Table 5).

292 The lowest C/N ratio was found in B+NM in spring and this value was significant different from  
293 the other five samples (B+M and B+NM in spring as well as NB, B+M and B+NM in autumn,  
294 all of which showing not significant differences). More specifically, the NB soils showed the

295 lowest C/N ratio ( $13.9 \pm 0.73$ ) in autumn and the highest values ( $22.1 \pm 2.23$ ) in spring. In autumn,  
296 the maximum value of C/N ratio ( $16.3 \pm 1.38$ ) was measured in B+M soils, while in spring the  
297 minimum C/N ratio ( $12.5 \pm 2.33$ ) was found in B+NM plots (Table 5). From these changes, a  
298 large reduction in C/N ratio (by 37%) was estimated in NB plots and an increase was calculated  
299 in both burned soils, more noticeable in B+NM plots (+20% against a 4% in B+M soil) (Table  
300 5).

301

### 302 *3.3. Differences among treatments and temporal changes in microbiological properties of soils*

303

304 Both in spring and autumn, the B+M soils generally showed the highest values in all the  
305 microbiological properties in comparison to other treatments, except for DHA and BSR (in  
306 autumn); the highest DHA and BSR in autumn were detected in NB and B+NM soils as well as  
307 in B+NM, respectively. Moreover, the differences in the enzymatic activity between B+M soils  
308 and the other treatments were significant for BGA, and UA in both seasons **CONTROLAR**  
309 **Acid-PA y GPRS** and MC in autumn.

310 Most of the surveyed microbiological properties attained the lowest values in the NB plots (e.g.,  
311 BGA and BSR in spring and autumn, Acid-PA and GPRS in autumn as well as DHA and MC in  
312 spring) (Table 6).

313 The soils sampled in B+NM plots showed the lowest UA (in both season), BGA, Acid-PA and  
314 GPRS (in spring) and MC (in autumn), whereas DHA and BSR were the highest among the  
315 treatments in spring and in autumn, respectively. Compared to the control soils, the differences  
316 are significant only for Acid-PA **CONTROLAR GPRS** in both seasons, UA in spring as well as  
317 GPRS and MC in autumn (Table 6).

318

319 **3.4. Correlations among physical, chemical and microbiological soil properties**

320

321 The Pearson's correlation analysis among the physical, chemical and microbiological soil  
322 properties surveyed in the experimental site showed interesting correlations (Table 7). As regards  
323 soil texture, the clay content was negatively correlated with silt ( $r = -0.93$ ) and sand ( $r = -0.63$ )  
324 contents. The soil pH was significantly linked with glomalin ( $r = -0.48$ ), while EC showed a  
325 higher correlation with chemical (OM, OC and TN contents,  $r > 0.77$ ) than with microbiological  
326 properties (BGA and UA). The highest correlations ( $r > 0.98$ ) were found among OC, OM and  
327 TN contents of the soils. As expected, a noticeable and significant  $r (> 0.55)$  was found between  
328 the C/N ratio and the OC and TN contents. Microbiological soil properties, except DHA, showed  
329 high significant correlations with several physical and chemical parameters (Table 7). BGA was  
330 the enzymatic activity that showed the greatest number of positive correlations ( $r > 0.49$ ) with  
331 physical and chemical soil properties, but it was negatively correlated with the clay content ( $r = -$   
332  $0.68$ ). Moreover, BGA, UA and Acid-PA were positively correlated each other with a minimum  
333  $r$  of  $0.61$  between BGA and UA and a maximum  $r$  of  $0.77$  between BGA and Acid-PA. GPRS  
334 was positively correlated with OM and OC contents ( $r = 0.47$  for both) and with BGA and  
335 negatively correlated with DHA ( $r = -0.53$ ). A correlation coefficient of  $0.54$  was found between  
336 the BSR and TN, OC and OM contents, while the MC was only positively correlated with GPRS  
337 ( $r = 0.50$ ) (Table 7).

338

339 **4. Discussions**

340

341 Studying the incidence of post-fire management actions on soil properties and the related  
342 changes is very important to identify the magnitude of these effects and plan possible  
343 countermeasures against soil degradation, but, due to the number and complexity of these effects,  
344 very little background is available. Therefore, it is necessary to select a set of soil parameters  
345 about physical, chemical and biological soils properties, which suitably reflect its status and the  
346 functions that need evaluating (Muñoz-Rojas et al., 2016). The extent of post-fire changes in  
347 some soil properties, directly attributed to heating, is usually related to burn severity (Mataix-  
348 Solera et al., 2009). Many authors have found changes in the soil quality and organic matter  
349 content in these soils (González-Pérez et al., 2004), increases in soil pH (Ulery et al., 1993;  
350 Mataix-Solera et al., 2002), decay of soil structure and thus the stability of aggregates, formation  
351 of hydrophobic films on soil aggregates (DeBano 2000), changes in the nutrient availability and  
352 water retention (Certini, 2005) and modifications of the enzymatic activities (Mataix-Solera et  
353 al., 2009). As regards the latter, since enzymatic activities have an important role in catalyzing  
354 biological reactions, there is a particular need of information about reaction rates related to  
355 production of essential elements in biogeochemical cycles (Mataix-Solera et al., 2009).

356 This study has explored the effects of straw mulching application immediately after wildfire on  
357 meaningful of some physical, chemical and microbiological soil properties in comparison to  
358 unthreatened and control plots, with particular regard on microbial activity, previously not  
359 enough investigated in Mediterranean environments (D'Ascoli et al., 2005; Pourreza et al., 2014;  
360 Rincón and Pueyo, 2010) and especially in forests. More specifically, it has been investigated  
361 whether the soil treatments with straw mulching could be beneficial for soil quality in terms of  
362 changes in physico-chemical and microbiological soil properties at short-term after fire  
363 occurrence, in order to mitigate the well-known soil degradation induced by fire.



364

365 In the same environment and experimental plots than in this study, [Lucas-Borja et al. \(2019\)](#)  
366 observed noticeable variations in vegetation cover, dead matter and bare soils extent one year  
367 after fire for each experimental condition (non-burned, burned and non-treated as well as burned  
368 and mulched soils). According to [Lucas-Borja et al. \(2019\)](#) and in spring 2017, soil became bare  
369 for about 30% and a vegetation cover of 10% was detected in the B+NM experimental plots. The  
370 covers of vegetation and dead matter (coming from straw application) were about 25% and 55%,  
371 respectively, in B+M plots. In the following autumn, the extent of bare soils increased to 55%  
372 and vegetation cover increased to 25% in B+NM soils, whereas the latter parameter was 53%  
373 and dead matter cover was 22% in B+M soils. In addition, straw mulching was found to promote  
374 a higher water content and a lower temperature of soil, determining sunlight interception ([Lucas-  
375 Borja et al., 2019](#)). From these findings, we suspected that these changes in soil vegetation cover  
376 and soil microclimatic conditions, significantly may have altered the physico-chemical and  
377 microbiological soil properties during the two sampled periods. Moreover, since all the  
378 experimental plots were set up on sites characterizes with the same burn severity, changes in soil  
379 properties may be not attributed to burn severity.

380 It is well known that, of all the physical and chemical soils properties, the OM content is one of  
381 the most important quality indicators, given its influence on plant growth-related functions (e.g.,  
382 humidity being retained, reservoir and nutrient exchange) ([Muñoz-Rojas et al., 2016](#)) and also on  
383 the maintenance of productivity, biodiversity and other ecosystem services ([Lucas-Borja et al.,  
384 2016](#)). In this study and both in autumn and spring, OM showed the highest values in B+M soils;  
385 this parameter was significantly lower in B+NM and NB plots. Also, variations in TN contents  
386 were detected after wildfire and mulching, inducing significant increases in B+NM and mainly in

387 B+M plots. It is expected that soil treatments (in our case fire and mulching) modify the C/N  
388 ratios compared to the values recorded before the fire. The simultaneous changes in OM and TN  
389 significantly reduced the C/N ratio only in B+NM soils immediately after the wildfire, since the  
390 C/N ratio is related with OM decomposition and N mineralisation (Lucas-Borja et al., 2016).  
391 After one year of time span, all the experimental soils did not show significant differences in  
392 C/N ratio, although a slight increase was recorded in B+M plots. This is in accordance with  
393 previous studies in burned pine forests, indicating that, after the initial C/N drop caused by fire,  
394 and owing to new forms of recalcitrant N accumulation and to the volatilisation of C  
395 compounds immediately after fire (Carballas et al., 2009; Rodríguez et al., 2017), the C/N ratio  
396 recovers its pre-fire values (Jiménez-González et al., 2016). A higher C/N ratio for the hillslope  
397 stabilisation-treated plots (as in our B+M soils) indicates low activity and disintegration speed  
398 for OM as well as a lower degree of N mineralization, which may be due to a more recalcitrant  
399 chemical composition of litter and low litter quality (high C/N ratio) (Martín-Peinado et al.,  
400 2016).

401 Changes on soil texture of burned soils (decrease of clay and increase of silt in B+NM plots and  
402 increase of silt and decrease of sand in B+M) are quite expected after wildfire, as also detected in  
403 the same environment by Lucas-Borja et al. (2019). These changes must be monitored with  
404 caution, since a decrease of the finer fraction let us suspect that burned but not treated soils may  
405 be more prone to erosion compared to non-burned and burned but mulched soils.

406 Literature shows that soil pH and EC tend to rise after fire, although, in any case, these properties  
407 gradually return to the original pre-fire values due to the washout effect (Mataix-Solera et al.,  
408 2009; Muñoz-Rojas et al., 2016). In the study area, pH did not respond to the described pattern,  
409 since its changes among treatments is stable, probably due to the higher buffering capacity of

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410 carbonated soils (Certini, 2005; Mataix-Solera et al., 2009). Conversely, EC of burned soils  
411 evolved as predicted by literature, since, after sudden increases immediately after fire compared  
412 to NB soils, this parameter strongly decreased with a more noticeable effect in B+NM plots. This  
413 may be due to the effects of burning, which accumulated ash that contains C and other nutrients  
414 from burned forest fuel (Caon et al., 2014). Mulching should have smoothed the decreasing trend  
415 of EC, thanks to the progressive release of these compounds.

416 With regards to the monitoring of the soil microbiological properties, we noticed that both the  
417 quantity and activity of microorganisms grew, as respectively indicated by biomass carbon and  
418 basal soil respiration (the later being not statistical different) parameters increased in the burned  
419 soils compared to NB plots immediately after fire. Conversely, one year after fire, the microbial  
420 carbon decreased in B+NM soils. These microbiological effects detected in B+M soils compared  
421 to non-treated plots may be due to the accumulation of biodegradable plant material and the  
422 increase in exchangeable cations (Rodríguez et al., 2017), which continued until these  
423 mineralized materials had been consumed (Muñoz-Rojas et al., 2016), with a lower effect  
424 recorded in B+NM soils. The mulching treatment had a remarkable effect on all microbiological  
425 and enzymatic activities, except for dehydrogenase. This was due to the accumulation of OM and  
426 nutrients and their following decomposition in soil throughout one year. This result was further  
427 confirmed by the positive correlations among the basal soil respiration, OM and TN content  
428 shown by Pearson's correlation analysis.

429 Our results showed different trends of enzyme content of soil, depending on their function and  
430 nature. In more detail, the recovery, and even the increase, in acid-phosphatase activity in B+M  
431 soils could be explained by its close relationship, stronger compared to the other enzymes, with  
432 the progressive restore of plant cover, with roots being the main resource (López-Poma and

433 [Bautista, 2014](#)). The lack of variation of dehydrogenase activity observed in the studied soils,  
434 beside the lack of response to the post-fire treatment with straw mulching, and their absence of  
435 relationships with the most of physical and chemical soil parameters, complies with other studies  
436 conducted in Mediterranean areas, showing the lack of sensitivity of dehydrogenase activity to  
437 seasonality and site effects than management practices. This could be related with the fact that  
438 dehydrogenases are not active as extracellular enzymes in soil, thus presenting a different pattern  
439 compared to the extracellular soil enzymes, that is,  $\beta$ -glucosidase, urease and acid-phosphatase  
440 (Blonska et al., 2017). The urease and  $\beta$ -glucosidase activities were greater in B+M soils both in  
441 spring and autumn compared to both NB and B+NM plots. The soil response of urease may be  
442 related with the greater accumulation of nitrogen due to the straw application, as indicated by the  
443 fair positive correlation with TN content shown by the Pearson's correlation analysis. The  
444 evolution of  $\beta$ -glucosidasae is related with OM decomposition velocity and with energy released  
445 by soil microorganisms, as indicated by the positive high correlation with the OC. The  
446 **progressive temporal changes among** the analyzed soil conditions suggests that mulching soils  
447 with straw could promote bacterial development, but the dehydrogenase activity could behave  
448 quite differently from the other enzymes.

449 Moreover, the glomalin content as a result of arbuscular micorrizal fungi is an indicator of C and  
450 N storage, which plays a key role in aggregate stability and water repellence of soils ([Lozano et](#)  
451 [al., 2016](#)). Although very few studies have explained its temporal evolution and response to post-  
452 fire mulching, some authors have demonstrated the glomalin sensitivity to fire, even at low  
453 temperatures ([Lozano et al., 2016](#)). For instance, [Rivas et al. \(2016\)](#) showed the GPRS level  
454 recovery four years after fire due to species' rapid root colonization that symbiosis undertakes  
455 with arbuscular micorrizal fungi. [Sansano \(2016\)](#) confirmed the negative influence of cutting

456 timber in the short term after fire. This study has demonstrated that glomalin content quickly  
457 recovers after fire, especially in B+M soils (presumably thanks to the higher OM content and  
458 C/N ratio), in accordance with Sansano (2016). Finally, the Canonical Analysis of Principal  
459 Coordinates has clearly discriminated NB, B+NM and B+M soil samples in terms of their  
460 physical, chemical and microbiological soils properties. This suggests that, one year after  
461 wildfire, burned soils not subject to any treatment present different physico-chemical and  
462 microbiological soil properties compared to soils treated with straw mulching and to control  
463 plots.

464

## 465 5. Conclusions

466

467 In order to better understand the effects of an important post-fire soil restoration technique, such  
468 as mulching with straw, on soil functionality of the Mediterranean forests, particularly prone to  
469 intense erosion and degradation, this study has evaluated the seasonal changes in the physical,  
470 chemical and microbiological properties of burned (treated with straw mulching or not) soils  
471 compared to non-burned plots throughout one year after a wildfire in a *Pinus halepensis* M.  
472 forest. Differentiated physical, chemical and microbiological properties between non-burned,  
473 burned and non-treated, and burned and straw-mulched soils were confirmed by the Canonical  
474 Analysis of Principal Coordinates. The results of enzyme activity monitoring in the soils have  
475 confirmed the working hypothesis that the straw mulching enhances soil functionality on the  
476 short-term. As a matter of fact, compared to non-burned soils, both in autumn and spring, in  
477 burned and straw-mulched soils the organic matter content increased. The C/N ratio, after a  
478 decrease immediately after fire in burned and non-mulched plots, did not show significant

479 differences among the other treatments. The values of pH were stable, but a marked reduction in  
480 the electric conductivity was noticed after one year, especially in burned soils both treated with  
481 straw-mulching and without mulching. The quantity and activity of soil microorganisms grew (as  
482 shown by the increase of biomass carbon and basal soil respiration) in the burned and treated  
483 soils. Moreover, after the mulching treatment all microbiological and enzymatic activities  
484 improved, except for dehydrogenase activity. The glomalin content quickly recovered after fire,  
485 indicating a higher carbon and nitrogen storage of mulched soils. One year after fire, burned soils  
486 not subject to any treatment were in an intermediate stage between the characteristics they had  
487 immediately after fire and those of soils treated with straw mulching.

488 Overall, the results of this study, indicating the positive effects of straw application in areas  
489 burned by high-severity fire on soil functionality, support forest managers and policy makers in  
490 selecting the proper management actions against soil erosion and degradation in the delicate  
491 environmental ecosystems of Mediterranean pine forests.

492

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494

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