

**NATURAL RADIOACTIVITY OF THE CRYSTALLINE BASEMENT
ROCKS OF THE PELORITANI MOUNTAINS (NORTH-EASTERN
SICILY, ITALY): MEASUREMENTS AND RADIOLOGICAL
HAZARD**

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2 **THE PELORITANI MOUNTAINS (NORTH-EASTERN SICILY, ITALY):**
3 **MEASUREMENTS AND RADIOLOGICAL HAZARD**

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20 Running Title: Natural radiation from NE Sicily rocks
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74 Crystalline rocks can produce dangerous radiation levels on the basis of their content in radioisotopes. Here, we report radiological data
75 from ten metamorphic and igneous rock samples collected from the crystalline basement of the Peloritani Mountains (Southern Italy). In
76 order to evaluate the radiological properties of these rocks, the gamma radiation and the radon emanation have been measured. Moreover,
77 since some of these rocks are employed as building materials, we assess the potential hazard for population connected to their use. Gamma
78 spectroscopy was used to measure the ^{226}Ra , ^{232}Th and ^{40}K activity concentration whereas the radon emanation was investigated by using a
79 RAD 7 detector. The results show ^{226}Ra , ^{232}Th and ^{40}K activity concentration values ranging from $(17 \pm 4) \text{ Bq kg}^{-1}$ to $(56 \pm 8) \text{ Bq kg}^{-1}$, (14 ± 3)
80 Bq kg^{-1} to $(77 \pm 14) \text{ Bq kg}^{-1}$ and $(167 \pm 84) \text{ Bq kg}^{-1}$ to $(1760 \pm 242) \text{ Bq kg}^{-1}$ respectively. Values of the annual effective dose equivalent (AEDE)
81 outdoor range from 0.035 to 0.152 mSv y^{-1} whereas the gamma index is in the range of $0.22 - 0.98$. The ^{222}Rn emanation coefficient and the
82 ^{222}Rn surface exhalation rate vary from $(0.63 \pm 0.3) \%$ to $(8.27 \pm 1.6) \%$ and from $(0.12 \pm 0.03) \text{ Bq m}^{-2} \text{ h}^{-1}$ to $(2.75 \pm 0.17) \text{ Bq m}^{-2} \text{ h}^{-1}$
83 respectively. The indoor radon derived from the building use of these rocks induces an approximate contribution to the annual effective dose
84 ranging from 8 to $176 \mu\text{Sv y}^{-1}$. All the obtained results suggest that the crystalline rocks from the Peloritani Mountains are not harmful for
85 the residential population, even though they induce annual effective doses due to terrestrial gamma radiation above the worldwide average
86 values. Moreover, their use as building materials does not produce significant health hazards connected to the indoor radon exposure.

87

88 INTRODUCTION

89 All living organisms on Earth, including the world population, are constantly exposed to ionizing
90 radiation coming from natural sources. Their effects to organs and tissues, if the radiation levels are
91 very high, can produce serious consequences on health. Natural radiation results from terrestrial
92 radionuclides occurring in the Earth's crust and mantle (e.g. ^{235}U , ^{238}U , ^{232}Th , ^{40}K and ^{87}Rb) and
93 from cosmic rays coming from the outer space and from the Sun. Two components contribute to the
94 exposure from natural radiation: i) the external exposure, caused by gamma radiation induced from
95 terrestrial radionuclides and by cosmic radiation and cosmogenic radionuclides, and: ii) the internal
96 exposure, related to inhalation and ingestion. According to UNSCEAR⁽¹⁾, the annual global average
97 effective dose received by an individual due to natural radiation results to be 2.4 mSv .

98 Among all the radionuclides responsible for the natural radiation exposure, radon and its daughters
99 (i.e. polonium and bismuth) are considered to provide the largest contribution⁽¹⁾. Radon is an
100 odorless inert noble gas and the main isotopes are: ^{222}Rn (half-life of 3.8 days), which occurs as
101 intermediate product in the ^{238}U decay series and it is generated by the ^{226}Ra α decay; and ^{220}Rn
102 (half-life of 55 s), that belongs to the decay serie for ^{232}Th and it is produced by the α decay of
103 ^{224}Ra . A strong relationship between indoor radon exposure and the increase of the risk of lung
104 cancer exists⁽²⁻⁶⁾. According to WHO⁽⁷⁾, after tobacco smoking, radon is considered the second
105 cause of lung cancer in the general population. Radon-related health hazards are connected to
106 indoor radon deriving from soils and from building materials and to radon dissolved in domestic
107 and drinking waters. Gamma rays from terrestrial radionuclides provide the second largest
108 contribution to the annual effective dose⁽¹⁾. Gamma radiation originates from ^{40}K and from
109 radionuclides occurring in the ^{238}U and ^{232}Th series.

110 The Peloritani Mountains, characterized by the presence of crystalline rocks (granites and low to
111 high-grade metamorphites), constitute the north-eastern portion of Sicily (Southern Apennines,
112 Italy), an area interested by a complicated geodynamic setting where subduction, transtensive
113 deformation and active extension processes coexist⁽⁸⁻¹⁴⁾. It is known that uranium and thorium occur
114 in low to high concentrations in unmetamorphosed and metamorphosed felsic igneous rocks,

115 because during magmatic processes, these highly incompatible elements tend to accumulate in the
116 residual Si-rich magma melt. Moreover, U and Th may also occur at high concentrations in meta-
117 sediments, especially if rocks interact with metasomatic and/or hydrothermal fluids at low to
118 medium temperatures. Consequently, metamorphic and felsic igneous rocks could represent
119 dangerous radiation sources, and all the outcropping areas, such as the Peloritani Mountains, can be
120 subjected to very high values of annual effective doses from natural radiation.

121 Taking into account these consideration, here we intend to assess the environmental impact and the
122 radiological hazards for inhabitants, connected to the natural occurrence of the crystalline rocks of
123 the Peloritani Mountains. Moreover, since these rocks are extensively used as building materials,
124 the potential hazards due to their use in construction was also evaluated. To reach this aim, a
125 multidisciplinary approach including gamma ray and alpha spectroscopy was used to constrain the
126 radiological properties in terms of gamma radiation and radon emanation.

127

128 GEOLOGICAL SETTING

129 The Peloritani Mountains represent the southernmost portion of the Calabria-Peloritani Orogen
130 (CPO) and constitute the north-eastern edge of Sicily (Figure 1). The CPO is composed of oceanic
131 units and a series of continental units including a metamorphic basement and often, a Meso-
132 Cenozoic sedimentary cover. The CPO is divided, from a physiographic point of view, in a northern
133 sector including Catena Costiera and Sila Massif and a southern sector represented by Serre and
134 Aspromonte Massifs in Calabria and Peloritani Mountains in Sicily⁽¹⁵⁾. In Sicily, the Peloritani Belt
135 overthrusts the Alpine Tethys units (i.e. Sicilide Units of Ogniben⁽¹⁶⁾) deformed during the
136 accretionary and collisional processes connected with the subduction and closure of the Alpine
137 Tethys realm.

138 The Peloritani Mountains include different continental units forming a south-verging nappe system,
139 which are characterized by distinct tectono-metamorphic portions divided into an upper and a lower
140 complex⁽¹⁵⁾. The upper complex outcrops in the northern sector of the Peloritani Mountains (Figure
141 1) and is composed of the Aspromonte and the Mandanici metamorphic Units, which are separated
142 by a regional thrust (the Aspromonte basal thrust⁽¹⁷⁾). The Aspromonte Unit represents the highest
143 tectono-metamorphic unit of the Peloritani Belt and is mainly composed of Variscan fine-grained
144 paragneiss and orthogneiss, with minor amphibolite gneisses and marbles. Late Variscan igneous
145 rocks (~290 Ma⁽¹⁸⁾) intruded the high-grade metamorphic rocks of the Aspromonte Unit, generating
146 dykes and small plutons (e.g. the stock of Capo Rasocolmo). The Mandanici Unit is composed of
147 medium to low-grade metamorphic rocks such as phyllites with minor metamorphosed volcanic and

148 carbonate rocks, which were interested by a greenschist to lower amphibolite Variscan
149 metamorphism.

150 The lower complex outcrops in the southern sector of the Peloritani Mountains (Figure 1), and it is
151 made up by three similar tectonic Units (St. Marco d'Alunzio, Longi-Taormina Unit and Cape St.
152 Andrea Units) including a Variscan crystalline basement overlaid by a Meso-Cenozoic sedimentary
153 cover. The metamorphic rocks of the lower complex are composed of metamorphosed fine-grained
154 sedimentary rocks (i.e. metapelites), sometimes interbedded by felsic porphyroids and meta-
155 volcanic rocks with mafic to intermediate compositions⁽¹⁹⁻²³⁾.

156 The metamorphic and plutonic basement rocks are locally covered by Meso-Cenozoic sediments,
157 unconformably overlaid by syn- to post-orogenic sedimentary sequences. The top of these
158 sedimentary sequences consists of Pleistocene terrigenous deposits (called "Sabbie and ghiaie of
159 Messina" formation) outcropping along the Tyrrhenian and the Ionian margins of the Peloritani
160 Belt. They represent the remnant of old fluvial and deltaic systems generated in response to the
161 uplift stage occurred in North-Eastern Sicily during the Quaternary⁽²⁴⁾. They are composed of
162 crystalline sands and gravels primarily deriving from erosion and dismantling of three different type
163 of rocks: Late Variscan plutonites, orthogneisses and paragneisses, all belonging to the Aspromonte
164 Unit.

166 MATERIALS AND METHODS

167 **Samples**

168 In this study, ten samples of crystalline rocks from the CPO basement outcropping in North-eastern
169 Sicily were collected. Locations and coordinates of the sampled rocks are listed and plotted in Table
170 1 and Figure 1, respectively. The collected samples include rocks from:

- 171 1. the Aspromonte Unit (ASP1: Late Variscan monzogranite; ASP2: orthogneiss; ASP3:
172 paragneiss; ASP4: amphibolite gneiss; ASP5: marble);
- 173 2. the Mandanici Unit (MAN1: phyllite);
- 174 3. the St. Marco d'Alunzio Unit (SMU1: metapelite);
- 175 4. the Longi-Taormina Unit (LTU1: metapelite; LTU2: felsic porphyroid; LTU3: intermediate
176 meta-volcanite).

177 ASP2, ASP3 and ASP5 samples are currently mined and used in construction. ASP4, MAN1,
178 SMU1, LTU1, LTU2 and LTU3 samples are not used as building materials, and therefore in this
179 paper we examine the possible effects deriving from their potential use. Furthermore, ASP1, ASP2
180 and ASP3 compositions are representative of the "Sabbie and ghiaie of Messina" formation, which
181 is extracted and used as inert material in bulk amounts (e.g. concrete).

182

183 Sample preparation

184 All samples were carefully cleaned in order to remove the macroscopic altered portions. Then, each
185 sample was crushed in an agate mill, and sieved in order to obtain a particle size less than 2 mm.
186 After that, in agreement with the procedure adopted in Lee et al.⁽²⁵⁾ and Sabatino et al.⁽²⁶⁾, samples
187 were dried in an oven for 24 h at a temperature of 105°C, and then cooled in a desiccator to constant
188 weight. Subsequently, samples were sealed in an airtight containers for more than four weeks; after
189 that, the secular equilibrium between ²²⁶Ra and their daughters was attained and samples were ready
190 to be analysed for gamma ray spectroscopy at the laboratories of the ARPACal Reggio Calabria
191 Department and for the radon emanation at the laboratories of the MIFT Department of the Messina
192 University. The radon emanation was measured by means of a homemade airtight container
193 connected to a RAD 7 spectrometer in terms of: a) emanation coefficient (i.e. emanation fraction,
194 emanation factor), representing the number of radon atoms released from the solid matrix with
195 respect the number of radon atoms generated by radium decay⁽²⁷⁾; b) exhalation rate, which is the
196 amount of radon released per unit time.

197

198 Gamma spectroscopy

199 Samples analyzed for the gamma spectrometry were packed in polyethylene plastic vial of 20 mL of
200 volume, in order to obtain a homogeneous geometry around the detector. The net weights of
201 samples were measured before the analysis.

202 Samples were counted for 70000 s and analysed to determine the ²²⁶Ra, ²³²Th and ⁴⁰K activity
203 concentration. The ²²⁶Ra activity concentration was determined through the 295.21 keV and 351.92
204 keV ²¹⁴Pb and 1120.29 keV ²¹⁴Bi γ -lines, whereas the 911.21 keV and 968.97 keV ²²⁸Ac γ -lines,
205 were used to calculate the ²³²Th activity concentration. For ⁴⁰K, the evaluation was performed from
206 its γ -lines at 1460.8 keV. The setup was composed by two Ortec HPGe detectors and integrated
207 digital electronics. The first detector is a negative biased detector (GMX) with FWHM of 1.94 keV,
208 peak to Compton ratio of 65:1 and relative efficiency of 37.5 % at 1.33 MeV (⁶⁰Co). The second
209 one is a positive biased detector (GEM) with FWHM of 1.85 keV, peak to Compton ratio of 64:1
210 and relative efficiency of 40 % at 1.33 MeV (⁶⁰Co). Lead shields with copper and tin lining were
211 used to shield the detectors from environmental background. The efficiency and the energy
212 calibration were performed by using suitable standards. In particular, the Eckert and Ziegler Nuclitec
213 GmgH traceable multinuclide radioactive standards (number AK 5901) were used; they cover the
214 energy range from 59.54 keV to 836 keV, and they were customized to clone the identical
215 geometries of samples in a water-equivalent epoxy resin matrix. Moreover, for the efficiency

216 transfer factors calculations with respect to the vial sample geometry, the ANGLE 4 code was
217 employed (Angle4 Software Home Page⁽²⁸⁾).

218 For data acquisition and analysis, the Gamma Vision (Ortec) software was used (see Caridi et al.⁽²⁹⁾
219 for further information). Key information such as energy, half-life, etc. about the investigated
220 radionuclides, are contained in an appropriate library, which was used to identify them in the
221 spectrum and then, to perform activity calculations and corrections.

222 The activity concentration of the investigated radioisotopes was calculated using the following
223 formula:

$$224 \quad C = \frac{N_E}{\varepsilon_E \gamma_d M t} \quad (1)$$

225 where N_E is the net area of the radioisotope photopeak, ε_E and γ_d indicate the efficiency and yield
226 of the photopeak, M represents the mass of the sample (g) and t indicates the live time (s). The
227 correction for the self-absorption effect was performed through the Gamma Vision software itself.

228 The measurement result uncertainty, coverage factor $k=2$, was determined considering the
229 uncertainty of: i) the counting estimation; ii) the calibration source; iii) the efficiency calibration;
230 iv) the background subtraction; v) the γ -branching ratio⁽²⁹⁾.

231

232 **Outdoor radiological hazard due to gamma radiation: the absorbed dose rate and the annual** 233 **effective dose equivalent (AEDE)**

234 For the absorbed dose rate (D) determination, conversion coefficients were used as follows⁽¹⁾:

$$235 \quad D = 0.462C_{Ra} + 0.604C_{Th} + 0.0417C_K \quad (2)$$

236 where C_{Ra} , C_{Th} and C_K are the activity concentrations (Bq kg^{-1}) of ^{226}Ra , ^{232}Th and ^{40}K in the
237 analysed samples, respectively. The absorbed dose rate is expressed in nGy h^{-1} .

238 The annual effective dose equivalent outdoor received by an individual who stays over an area
239 formed of the studied rocks ($AEDE_{outdoor}$, mSv y^{-1}), was estimated using the adsorbed dose rate by
240 using a conversion factor of 0.7 Sv Gy^{-1} from adsorbed dose to effective dose, and an outdoor
241 occupancy factor of 0.2 ⁽¹⁾:

$$242 \quad AEDE_{outdoor} = D \cdot 8760 \text{ h} \cdot 0.2 \cdot 0.7 \cdot 10^{-6} \quad (3)$$

243

244 **Exposure to gamma radiation originating from building materials: the Gamma index**

245 The exposure to gamma radiation originating from radionuclides in building materials was
246 evaluated through the “gamma index”, also known as “concentration index”^(30, 31) (I):

$$247 \quad I = \frac{C_{Ra}}{300 \text{ Bq kg}^{-1}} + \frac{C_{Th}}{200 \text{ Bq kg}^{-1}} + \frac{C_K}{3000 \text{ Bq kg}^{-1}} \quad (4)$$

248 where C_{Ra} , C_{Th} and C_K are the activity concentrations expressed in Bq kg⁻¹ of ²²⁶Ra, ²³²Th and ⁴⁰K.
 249 The gamma index is generally used only as a screening tool for identifying materials that may
 250 induce a significant amount of gamma radiation⁽³⁰⁾.

251

252 **Radon emanation measurements**

253 The ²²²Rn emanation coefficient was evaluated by enclosing 0.8 L of sample material in an airtight
 254 container whose volume is equal to 8 L, causing the growth of the radon activity concentration in
 255 the air volume. The experimental setup is similar to that reported in IAEA⁽³²⁾ and in Arabi et al.^{(33,}
 256 ³⁴⁾. The container was connected to a dryer unit filled with calcium sulphate (CaSO₄) and to an
 257 electronic radon detector (RAD 7, Durrige Co.) through a piping system forming a closed loop in
 258 which air circulates. Silicon sealing was installed along the piping system and along connections
 259 among the container and the pipes.

260 The RAD 7 Durrige was used to measure the radon activity concentration in the air volume. The
 261 RAD 7 detector (see the RAD 7 Manual⁽³⁵⁾) is a solid-state ion-implanted planar silicon alpha
 262 detector, which is located inside an internal cell of 0.7 L. The measurement of radon activity
 263 concentrations is achieved by using the activity of relative daughters (i.e. polonium isotopes). The
 264 internal cell, having a hemispherical shape, is coated on the inside with an electrical conductor
 265 charged by a high voltage power circuit able to develop a potential of 2000/2500 V with respect to
 266 the alpha detector. Radon nuclei decaying within the internal cell produce transformed nuclei (i.e.
 267 polonium) as positively charged ions. The electric field inside the cell, allows to drive the positive
 268 charged particles straight onto the detector, which is arranged at the centre of the hemisphere. The
 269 short-lived polonium nuclei decay upon the detector's active surface, and the emitted alpha particles
 270 having a 50% probability of entering the detector. The electrical signal produced whenever an alpha
 271 particle enter the detector is proportional in strength to the energy of the alpha particle. Different
 272 isotopes have different alpha energies, and produce different strength signals. Concerning our
 273 experiment, the "Sniff" mode (see the RAD 7 Manual⁽³⁵⁾) was adopted. With this setup, the RAD 7
 274 Durrige uses only the ²¹⁸Po alpha peak to determine the ²²²Rn concentration.

275 Samples were sealed in the airtight container for more than four weeks, in order to ensure the
 276 establishment of the secular equilibrium between ²²⁶Ra and ²²²Rn⁽³²⁾. After that, the activity
 277 concentration of ²²²Rn released from the sample into the air volume was measured by the RAD 7 by
 278 using a cycle of four hours in "Sniff" mode. Then, the ²²²Rn emanation coefficient (ε) was
 279 determined by the following formula⁽³²⁻³⁴⁾:

280

$$\varepsilon = \frac{VC_{Rn}}{MC_{Ra}} \quad (5)$$

281 where C_{Rn} (Bq m⁻³) is the measured radon activity concentration, C_{Ra} (Bq kg⁻¹) is the radium
 282 activity concentration, M (kg) is the mass of the sample and V (8.3×10^{-3} m³) is the effective volume
 283 of the system calculated by considering the volume of the container, of the sample, of the loop
 284 system and of the RAD 7 internal cell. Moreover, in order to avoid back diffusion effects, the
 285 effective volume of the system was >10 times larger than the pore volume of the samples^(36, 37).

286 The radon activity concentration in the container increases exponentially and tends to an
 287 equilibrium value. Knowing the equilibrium value, which corresponds to C_{Rn} (i.e. the activity
 288 concentration of ²²²Rn measured by the RAD 7 detector), it is possible to calculate the ²²²Rn
 289 exhalation rate, as follows^(32, 38):

$$290 \quad E = \frac{C_{Rn}(\lambda + \lambda l)V}{X} \quad (6)$$

291 where λ (7.56×10^{-3} h⁻¹) is the radon decay constant, λl is the container leakage rate (6.96×10^{-3} h⁻¹)
 292 ¹), V (8.3×10^{-3} m³) is the effective volume of the system and X is the sample's surface area (m²) or
 293 the mass of the sample (kg) depending on whether the exhalation rate is expressed as surface
 294 exhalation rate (E_s , Bq m⁻² h⁻¹) or mass exhalation rate (E_m , Bq kg⁻¹ h⁻¹), respectively.

295 The leakage from container (λl) was determined by continuously measuring the ²²²Rn concentration
 296 exhaled from one of the samples (i.e. the Mandanici phyllite) in the container at interval of 2 hours
 297 for 6 days. The series of the measured ²²²Rn concentrations were fitted to the equation expressing
 298 the radon activity growth as a function of time⁽³⁸⁾:

$$299 \quad C_t = C_0 e^{-t(\lambda + \lambda l)} + C_e (1 - e^{-t(\lambda + \lambda l)}) \quad (7)$$

300 where C_t is the ²²²Rn concentration at time t , C_0 and C_e are the ²²²Rn concentration at $t=0$ and at
 301 equilibrium, respectively, and λ is the radon decay constant. The experimental data fitted well with
 302 the model described, and λl of the experimental setup was obtained.

304 **Calculation of indoor radon concentration and indoor effective dose**

305 The indoor radon concentration depends on various factors, such as radon coming from soil beneath
 306 buildings and radon exhaled from building materials. It is possible to estimate the contribute to the
 307 indoor radon activity concentration in a room ($C_{iRn(bm)}$, Bq m⁻³), originating from radon exhalation
 308 from building materials, knowing their surface exhalation rates by the following equation⁽³⁸⁻⁴⁰⁾:

$$309 \quad C_{iRn(bm)} = \left[\sum_{i=1}^n w_{si} \cdot E_{Si} \right] \frac{S}{V\lambda_v} \quad (8)$$

310 where E_{si} ($\text{Bqm}^{-2}\text{h}^{-1}$) is the surface exhalation rate of the building material, w_{si} is the fractional usage
 311 of the building material, S/V (m^{-1}) is the surface to volume ratio of the room and λ_v (h^{-1}) is the
 312 annual average ventilation rate.

313 In addition, the annual effective dose induced by indoor radon exhaled from building materials (H ,
 314 $\mu\text{Sv y}^{-1}$) can be calculated using an indoor occupancy factor of 0.8, a ^{222}Rn equilibrium factor of
 315 0.4, and a dose conversion factor for ^{222}Rn of $9 \text{ nSv} (\text{Bq h m}^{-3})^{-1}$ as follows⁽¹⁾:

$$316 \quad H = C i_{Rn(bm)} \cdot 8760 h \cdot 0.8 \cdot 0.4 \cdot 9 \cdot 10^{-3} \quad (9)$$

317

318 RESULTS AND DISCUSSION

319 ^{226}Ra , ^{232}Th and ^{40}K occurring in the Peloritani rocks and soils generate a natural gamma radiation
 320 that produce exposures to humans. The activity concentrations of these terrestrial radionuclides
 321 were measured and adopted to compute the absorbed dose rate and the annual effective dose
 322 equivalent outdoor. Moreover, to test if the use in building construction might be of concern, the
 323 gamma index (I) was calculated.

324 The radon emanation from rock samples was investigated by calculating the ^{222}Rn emanation
 325 coefficient and the ^{222}Rn surface and mass exhalation rate. The ^{222}Rn surface exhalation rate was
 326 used to perform a rough preliminary estimation of the contribution to the radon indoor, of radon
 327 originating from the use of the Peloritani crystalline rocks as building materials as well as the
 328 effective doses induced to residents.

329

330 **Outdoor gamma radiation**

331 The values of activity concentrations of ^{226}Ra , ^{232}Th and ^{40}K in the investigated samples are
 332 reported in Table 2. Experimental results range from $(17 \pm 4) \text{ Bq kg}^{-1}$ to $(56 \pm 8) \text{ Bq kg}^{-1}$, from $(14$
 333 $\pm 3) \text{ Bq kg}^{-1}$ to $(77 \pm 14) \text{ Bq kg}^{-1}$ and from $(167 \pm 84) \text{ Bq kg}^{-1}$ to $(1760 \pm 242) \text{ Bq kg}^{-1}$ for ^{226}Ra ,
 334 ^{232}Th and ^{40}K , respectively.

335 LTU2 and LTU3 samples show the highest activity concentrations of respectively ^{226}Ra ($56 \pm 8 \text{ Bq}$
 336 kg^{-1}) and ^{40}K ($1760 \pm 242 \text{ Bq kg}^{-1}$), whereas the highest specific activity of ^{232}Th was detected in
 337 ASP2, ASP3, MAN1 and LTU3 samples, which exhibit values higher than 60 Bq kg^{-1} .

338 On the other hand, the ASP5 sample displays the lowest experimental values for ^{226}Ra ($17 \pm 4 \text{ Bq}$
 339 kg^{-1}) and ^{232}Th ($14 \pm 3 \text{ Bq kg}^{-1}$), while the ASP4 sample exhibits the lowest ^{40}K value ($167 \pm 84 \text{ Bq}$
 340 kg^{-1}).

341 The activity concentrations of ^{226}Ra , ^{232}Th and ^{40}K in the crystalline basement of the Peloritani Belt
 342 were compared with literature data of other European Countries and with the average worldwide
 343 values (UNSCEAR⁽¹⁾ and references therein). The comparison, shown in Figure 2, highlights that

344 the Peloritani rocks exhibit values of ^{40}K activity concentration higher than those of both European
345 Countries and worldwide average, whereas ^{226}Ra and ^{232}Th specific activities result to be quite
346 similar.

347 The adsorbed dose rate (D) and the corresponding annual effective dose equivalent outdoor ($AEDE_{outdoor}$)
348 were calculated to estimate the radiological hazard effect due to terrestrial gamma
349 rays emitted from the analysed rocks. The results are reported in Table 2. The adsorbed dose rates
350 induced by the crystalline rocks are in the range of 28 - 123 nGy h⁻¹.

351 ASP2 and LTU2 samples show the highest $AEDE_{outdoor}$ value: 0.152 mSv y⁻¹. Values above 0.1
352 mSv y⁻¹, were calculated for ASP3, MAN1 and LTU3 samples. Conversely, the lowest values are
353 those of ASP4 and ASP5 samples, 0.035 mSv y⁻¹ and 0.074 mSv y⁻¹ respectively.

354 All values except that for the ASP4 sample are higher than the worldwide average value of 0.07
355 mSv y⁻¹(⁴¹). Despite this evidence, according to the Italian legislation(⁴¹) these values are not
356 hazardous for inhabitants who spend their lives over areas characterized by the presence of the
357 crystalline rocks studied in this paper.

358 However, those values should be referred to and individual who stays over the studied rocks. If soil
359 is present above the bedrock, the dose rate will be presumably lower than that for a rock with the
360 same natural activity concentration of ^{226}Ra , ^{232}Th and ^{40}K , since soils are less dense than rocks and
361 usually they contain a percentage of moisture.

362

363 **Gamma Index**

364 The gamma index (I) was calculated with the aim of evaluating the hazard connected to the
365 exposure to gamma rays emitted from the crystalline rocks when they are used in construction. The
366 gamma index acts as a screening tool in identifying materials that could produce an increase to the
367 annual effective dose higher than 1 mSv. According to EC(³¹), dose increases of 1 mSv y⁻¹ are
368 reached with I values of 1 if the material is used in bulk amounts, and with I values 6 if the material
369 is used as tiles, boards or other restricted uses. Building materials should be exempted from
370 restrictions when their gamma radiation increases the annual effective doses up to 1 mSv y⁻¹, as
371 referred in Article 75 reported in Euratom Directive (³⁰).

372 All the analysed samples display I values ≤ 1 (Table 2); therefore, the building use is allowed.
373 However, ASP2, ASP3, MAN1, LTU2 and LTU3 samples show gamma index values between 0.76
374 and 0.98, which are slightly below the reference limit. Then, controls and additional measurements
375 for these lithologies are suggested.

376 Concerning the “Sabbie and ghiaie of Messina”, which are broadly used in bulk amounts (as inert in
377 concrete), the mineralogical content of these deposits abruptly varies from an area to another one in

378 relationship with the prevailing lithology from which the sediments were locally eroded. For this
379 reason, and since radiological and mineralogical properties of sediments reflect those of the eroded
380 parent rocks, we decided not to collect a sample of “Sabbie and ghiaie of Messina”, but to analyse
381 only their parent rocks (ASP1, ASP2 and ASP3 samples). Since the parent rock samples show I
382 values in the range of 0.62 - 0.98, we can assume that additional measurements are required for
383 safety use of the “Sabbie and ghiaie of Messina”.

384

385 **Natural radioactivity due to radon**

386 The ^{222}Rn emanation coefficients (ε) of the studied samples, calculated through the activity
387 concentration of ^{222}Rn released into the air volume of the sealed container, are reported in Table 3.
388 The emanation coefficient depends on various factors: radium distribution, grain size, grain shape,
389 temperature, moisture, specific surface and mineralogy^(27, 32, 42, 43). Also alteration and weathering
390 processes can affect the emanation coefficient value^(44, 45). The ^{222}Rn emanation coefficient of the
391 studied samples vary from $(0.63 \pm 0.3) \%$ to $(8.27 \pm 1.6) \%$ (Table 3). These values are comparable
392 with representative literature data reporting emanation coefficients from soils and rocks in the range
393 of 0.1 - 40 % (Sakoda et al.⁽²⁷⁾ and references therein). Since experimental conditions (i.e.
394 temperature, moisture, radium distribution, grain size and shape) are similar for all samples, the
395 only factors responsible for the different values of emanation coefficient among the crystalline
396 rocks are the mineralogical content and the alteration degree.

397 The Late Variscan monzogranite (ASP1 sample) shows an emanation coefficient value of $(2.44 \pm$
398 $0.9) \%$, which is comparable with values of unaltered granites (around 1 - 2 %^(34, 45)) since
399 weathered and altered granites can reach values higher than 20 %⁽⁴⁵⁾.

400 A wide record of data is reported in the literature for gneisses^(34, 46), which commonly exhibit ε
401 values varying from 1 to 14%. The results of this study fit well in that range, as orthogneiss (ASP2
402 sample) and paragneiss (ASP3 sample) display emanation coefficients of $(8.27 \pm 1.6) \%$ and $(2.30 \pm$
403 $0.7) \%$, respectively.

404 The ε values of $(5.81 \pm 1.3) \%$ measured for the phyllite of the Mandanic Unit (MAN1 sample) is
405 quite similar to literature data for phyllitic metamorphic rocks $(7.7 \%)^{(25)}$.

406 The Aspromonte marble (ASP5 sample) was collected at Cape Tindari, where the active tectonic
407 lineaments of the Tindari Fault System allow the upward rise of mantle and crustal fluids which can
408 easily alter the hosting rocks⁽⁴⁷⁾. The authors detected an intense rotten-eggs smell, which is a
409 typical feature of sulfurous gases and minerals, during sampling activities and during rock crushing
410 in laboratory. It could be an evidence that marbles suffered a low degree secondary alteration;
411 although no secondary minerals (e.g. sulfur-bearing minerals) were detected in the XRPD pattern.

412 Therefore, the secondary alteration processes supposedly affecting marbles might have modified the
413 pristine radon emanation power of the ASP5 sample as weathering and alteration usually increase
414 the emanation coefficient of rocks^(44, 45). Nevertheless, the ASP5 sample exhibits a ²²²Rn emanation
415 coefficient of $(5.99 \pm 1.9) \%$, which lies inside the range for carbonate rocks provided by previous
416 studied (limestones: 1.6 - 2.2 % and 6.8 - 13.2 %; Barretto⁽⁴⁶⁾ and Lee et al.⁽²⁵⁾, respectively;
417 marbles: 8.5 % and 9.54 - 16.42 %; El Afifi et al.⁽⁴⁸⁾ and Misdaq and Amghar⁽⁴⁹⁾, respectively).

418 The radon emanation coefficients calculated in this work can be used to estimate the theoretical
419 ²²²Rn activity concentration (i.e. the chemical ²²²Rn background concentration) in soils formed from
420 the crystalline rocks studied here⁽⁵⁰⁾. Soil radon values higher than the chemical background,
421 generally highlight the presence of radon anomalies related to the occurrence of fault zones
422 allowing the upward migration of deep gases (e.g. CO₂, CH₄, He and Rn).

423 Table 3 summarizes the ²²²Rn surface (E_S) and mass (E_M) exhalation rates calculated for each
424 samples. E_S and E_M values were derived to be in the range of $(0.12 \pm 0.03) \text{ Bq m}^{-2} \text{ h}^{-1}$ - (2.75 ± 0.17)
425 $\text{Bq m}^{-2} \text{ h}^{-1}$ and $(2.47 \pm 0.69) \text{ mBq kg}^{-1} \text{ h}^{-1}$ - $(54.02 \pm 3.26) \text{ mBq kg}^{-1} \text{ h}^{-1}$ respectively.

426 The highest ²²²Rn surface exhalation rates $(2.75 \pm 0.17 \text{ Bq m}^{-2} \text{ h}^{-1}$ and $1.56 \pm 0.12 \text{ Bq m}^{-2} \text{ h}^{-1}$) were
427 measured for the orthogneiss (ASP2 sample) and the phyllite (MAN1 sample), respectively. These
428 values have been compared with those of a number of natural materials often used in construction in
429 Italy. The ²²²Rn surface exhalation rates of orthogneisses and phyllites from Peloritani Mountains
430 are similar to those of several volcanic products from the Roman comagmatic province (Central
431 Italy) such as the “Tufo Lionato” from Colli Albani district $(0.86 - 1.40 \text{ Bq m}^{-2} \text{ h}^{-1}$; Tuccimei et
432 al.⁽³⁷⁾), the “Peperino” from Colli Albani districts $(1.33 - 3.42 \text{ Bq m}^{-2} \text{ h}^{-1}$; Tuccimei et al.⁽³⁷⁾), the
433 “Lava” from Nemi $(2.23 - 2.48 \text{ Bq m}^{-2} \text{ h}^{-1}$; Tuccimei et al.⁽³⁷⁾), and the “Lapilli” from Colli Albani
434 districts $(2.41 - 4.01 \text{ Bq m}^{-2} \text{ h}^{-1}$; Tuccimei et al.⁽³⁷⁾). However, orthogneiss and phyllite surface
435 exhalation rates are much lower than those of building materials characterized by a high attitude to
436 release radon (e.g. “Black pozzolana” from Colli Albani district $37 \text{ Bq m}^{-2} \text{ h}^{-1}$, “Tufo rosso” from
437 Vico district $5.47 - 17.06 \text{ Bq m}^{-2} \text{ h}^{-1}$ and “Tufo giallo napoletano” from Campania Region $2.66 -$
438 $10.33 \text{ Bq m}^{-2} \text{ h}^{-1}$; Tuccimei et al.⁽³⁷⁾); then, we should not considered them as a severe menace to the
439 public health.

440 Concerning the Aspromonte marble (ASP5 sample), it display a ²²²Rn surface exhalation rate of
441 $(0.82 \pm 0.09) \text{ Bq m}^{-2} \text{ h}^{-1}$, which is higher than those of the most famous carbonate rocks used as
442 building materials in Italy (e.g. “Rosso Veronese” from Veneto Region $0.04 \text{ Bq m}^{-2} \text{ h}^{-1}$, and
443 “Travertino” from Tivoli $0.05 - 0.065 \text{ Bq m}^{-2} \text{ h}^{-1}$; Tuccimei et al.⁽³⁷⁾). However, marble samples
444 analysed in other studies showed surface exhalation rates rather similar to that calculated for the
445 ASP5 sample (e.g. $0.74 \text{ Bq m}^{-2} \text{ h}^{-1}$; Kumar et al.⁽⁵¹⁾).

446 The E_S value of $0.30 \pm 0.05 \text{ Bq m}^{-2} \text{ h}^{-1}$ measured for the Late Variscan monzogranite (ASP1 sample)
447 is in agreement with the radon exhalation data coming from coeval granitoid rocks intruding the
448 continental crust of Sardinia (0.11 to $0.91 \text{ Bq m}^{-2} \text{ h}^{-1}$; Dentoni et al.⁽³⁸⁾). Late Variscan plutonites of
449 Peloritani Mountains and Sardinia also show analogous ^{226}Ra activity concentrations ($18 \pm 4 \text{ Bq kg}^{-1}$
450 1 of the ASP1 sample vs 18 to 80 Bq kg^{-1} of the Sardinian rocks⁽³⁸⁾). For these reasons, as already
451 suggested by Dentoni et al.⁽³⁸⁾, we carefully believe that the intrusive rocks generated within the
452 European continental crust of Sardinia, Sicily and Calabria during the latest phase of the Variscan
453 Orogeny (hence they formed in the same geodynamic setting), are characterized by similar
454 radiological features.

455 Radium content exhibits a very poor correlation when plotted against radon emanation coefficient,
456 surface exhalation rate and mass exhalation rate (not shown). Several studies^(38, 52, 53) highlighted a
457 similar behaviour testifying that radon emission from rocks mainly depends on mineralogical and
458 physical characteristics of the solid matrix.

459

460 **Hazard from radon in indoor settings**

461 The indoor radon in dwellings, schools and workplaces mainly originates from soil gas infiltration.
462 The reference level to minimize health hazards due to indoor radon exposure is 300 Bq m^{-3}
463 (WHO⁽⁷⁾). Radon emanation from building materials is the second major source to indoor radon,
464 and considering a model building in a temperate climate, its contribution to the indoor radon
465 activity concentration is around 20% ⁽⁵⁴⁾.

466 As a result, determining the radon emanation from building materials might be very helpful to
467 estimate health risks connected with the indoor radon exposure. For this purpose, we performed a
468 preliminary evaluation of the contribution to the indoor radon deriving from the use of the
469 Peloritani crystalline rocks in construction.

470 The indoor radon concentration coming only from radon emanation from building materials ($Ci_{Rn(bm)}$)
471 was calculated by equation 8 by using the estimated surface exhalation rates for the
472 crystalline rocks, a surface/volume ratio $S/V = 1.6 \text{ m}^{-1}$ considering a room model with dimensions
473 of $(4 \times 5 \times 2.8) \text{ m}^3$ and a ventilation rate $\lambda_v = 0.63 \text{ h}^{-1}$ (UNSCEAR⁽¹⁾). Moreover, equation 9
474 assesses the human exposure to indoor radon through calculation of annual effective doses (H). The
475 calculated $Ci_{Rn(bm)}$ and H values are reported in Table 3.

476 The orthogneiss (ASP2 sample) shows a $Ci_{Rn(bm)}$ of 6.99 Bq m^{-3} that corresponds to an effective
477 dose of $176 \mu\text{Sv}$ per year. These values are the highest among the collected samples.

478 The phyllite (MAN1 sample) and the felsic porphyroid (LTU2 sample) exhibit indoor radon activity
479 concentrations higher than 3 Bq m^{-3} and effective doses induced to humans of $100 \mu\text{Sv y}^{-1}$ and 81
480 $\mu\text{Sv y}^{-1}$, respectively. In contrast, other samples show H values lower than $60 \mu\text{Sv y}^{-1}$.

481 Considering the indoor radon reference limit of 300 Bq m^{-3} and a 20 % contribution from building
482 materials, the analysed rocks should produce $Ci_{Rn(bm)}$ values not exceeding 60 Bq m^{-3} in order to
483 minimize health hazards. All the crystalline rocks studied in this paper show $Ci_{Rn(bm)}$ values well
484 below the 60 Bq m^{-3} threshold (Table 3); therefore, their use as building materials does not produce
485 significant health risks due to indoor radon exposure. Also the use of the “Sabbie and ghiaie of
486 Messina” formation seems to be not hazardous as their parent rocks (paragneisses, monzogranites
487 and orthogneisses) exhale radon at low rates.

488 The indoor radon concentrations originating from the use of the studied materials, and the
489 corresponding annual effective doses received by humans, represent only rough estimations as some
490 limitations have to be taken into account: i) in the calculation we considered walls, the ceiling and
491 the floor exclusively composed of the analysed material (surface fractional usage $w_{st} = 100\%$); ii)
492 the “Sabbie and ghiaie of Messina” formation is largely used in construction, but collecting a
493 representative sample was not possible due to the huge vertical and lateral variability of the
494 mineralogical content within this formation; for this reason we analysed only their parent rocks
495 (paragneisses, orthogneisses and monzogranites); iii) we assumed walls, the ceiling and the floor
496 made up by the studied sample in the same form as during the experiment, namely in granular form;
497 but we have to consider that both radon diffusion length and radon surface exhalation rate, on which
498 the indoor radon concentration depends, can vary according to material shape and fabric.

499 Therefore, the calculated $Ci_{Rn(bm)}$ and H values should be attributed only to extreme situations,
500 adopted to preliminarily assess the radon potential risk associated with the use in construction of the
501 crystalline rocks from Peloritani Mountains.

502 CONCLUDING REMARKS 503

504 This study demonstrates that the crystalline rocks outcropping in the Peloritani Mountains do not
505 induce hazardous annual effective doses due to gamma rays emitted by terrestrial radionuclides
506 (^{226}Ra , ^{232}Th and ^{40}K). Moreover, their usage as building materials is not harmful in terms of radon
507 exposure. In fact, these rocks seem not to produce a contribution to the indoor radon concentrations
508 that are able to cause the exceeding of the recommended reference limit of 300 Bq m^{-3} , representing
509 the maximum acceptable level to minimize health hazards.

510 Nevertheless, since North-eastern Sicily is affected by an intense seismic activity⁽⁵⁵⁾, and the
511 relationship between tectonic activity and anomalous emission of gases including radon was widely

512 demonstrated by several scientific studies⁽⁵⁶⁻⁶¹⁾, we cannot completely ruled out the possibility that
 513 along fault zones the upward migration of deep fluids produces high natural radiation levels due to
 514 radon.

515 Radon released from soils can easily enter buildings from floor and walls cracks and discontinuities;
 516 underground spaces and rooms located at ground floor are the most vulnerable places. For this
 517 reason, we believe that further radiological analyses should be performed in order to better
 518 constraint if there are specific areas where the resident population is exposed to dangerous radiation
 519 doses. For instance, a series of surveys involving the measurements of soil radon concentration
 520 along tectonic lineaments and the evaluation of indoor radon levels in houses, schools and public
 521 edifices could be useful.

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688689 **Tables**

690 Table 1. Information, location and coordinates (UTM WGS84 zone 33S reference system) of the collected samples.

Sample ID	Type of rock	Tectonic Unit	Locality	North	East
ASP1	Late Variscan monzogranite	Aspromonte	Tono	4237468	549973
ASP2	Orthogneiss	Aspromonte	Marmara	4235290	541451
ASP3	Paragneiss	Aspromonte	Dinnammare	4223617	540613
ASP4	Amphibolite gneiss	Aspromonte	Dinnammare	4223594	540971
ASP5	Marble	Aspromonte	Tindari	4220994	504626
MAN1	Phyllite	Mandanici	Percia Rovetti	4211439	516590
SMU1	Metapelite	St. Marco d'Alunzio	Frascianida	4207313	516633
LTU1	Metapelite	Longi-Taormina	P.Illa Pandolfo	4191929	524118
LTU2	Felsic porphyroid	Longi-Taormina	Castelmola	4190603	524346
LTU3	Intermediate meta-volcanite	Longi-Taormina	Castelmola	4190149	524643

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Table 2. Activity concentration of ^{226}Ra , ^{232}Th and ^{40}K detected in the samples coming from the metamorphic basement of the Peloritani Belt, along with the calculated values of absorbed dose rate, annual effective dose outdoor and gamma index.

Sample ID	^{226}Ra (Bq kg ⁻¹)	^{232}Th (Bq kg ⁻¹)	^{40}K (Bq kg ⁻¹)	Adsorbed dose rate, <i>D</i> (nGy h ⁻¹)	Annual effective dose outdoor, <i>AEDE_{outdoor}</i> (mSv y ⁻¹)	Gamma index, <i>I</i>
ASP1	18 ± 4	31 ± 9	1220 ± 176	78	0.096	0.62
ASP2	45 ± 6	77 ± 14	1345 ± 190	123	0.152	0.98
ASP3	40 ± 7	68 ± 17	1009 ± 163	102	0.125	0.81
ASP4	27 ± 5	15 ± 2	167 ± 84	28	0.035	0.22
ASP5	17 ± 4	14 ± 3	1060 ± 165	61	0.074	0.48
MAN1	40 ± 6	64 ± 15	1063 ± 168	101	0.124	0.81
SMU1	25 ± 4	36 ± 9	595 ± 100	58	0.071	0.46
LTU1	34 ± 6	39 ± 11	674 ± 127	67	0.083	0.53

LTU2	48 ± 7	47 ± 12	1760 ± 242	123	0.152	0.98
LTU3	56 ± 8	62 ± 14	785 ± 135	96	0.118	0.76

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Table 3. Emanation coefficients and surface exhalation rates for the metamorphic samples resulting from the radon emanation experiment. The potential indoor radon activity concentration and the annual effective dose were also estimated.

Sample ID	Emanation coefficient, ϵ (%)	Surface exhalation rate, E_S (Bq m ⁻² h ⁻¹)	Mass exhalation rate, E_M (mBq kg ⁻¹ h ⁻¹)	Indoor ²²² Rn activity concentration, $C_{Rn(bm)}$ (Bq m ⁻³)	Annual effective dose indoor, H (μSv y ⁻¹)
ASP1	2.44 ± 0.9	0.30 ± 0.05	6.38 ± 1.10	0.77	19
ASP2	8.27 ± 1.6	2.75 ± 0.17	54.02 ± 3.26	6.99	176
ASP3	2.30 ± 0.7	0.63 ± 0.08	13.33 ± 1.70	1.60	40
ASP4	0.63 ± 0.3	0.12 ± 0.03	2.47 ± 0.69	0.30	8
ASP5	5.99 ± 1.9	0.82 ± 0.09	14.80 ± 1.63	2.08	53
MAN1	5.81 ± 1.3	1.56 ± 0.12	33.73 ± 2.69	3.95	100
SMU1	2.80 ± 0.8	0.50 ± 0.07	10.15 ± 1.45	1.27	32
LTU1	2.51 ± 0.8	0.62 ± 0.08	12.39 ± 1.62	1.57	39
LTU2	3.72 ± 1.1	1.27 ± 0.19	25.91 ± 3.78	3.22	81
LTU3	1.68 ± 0.4	0.72 ± 0.09	13.69 ± 1.62	1.83	46

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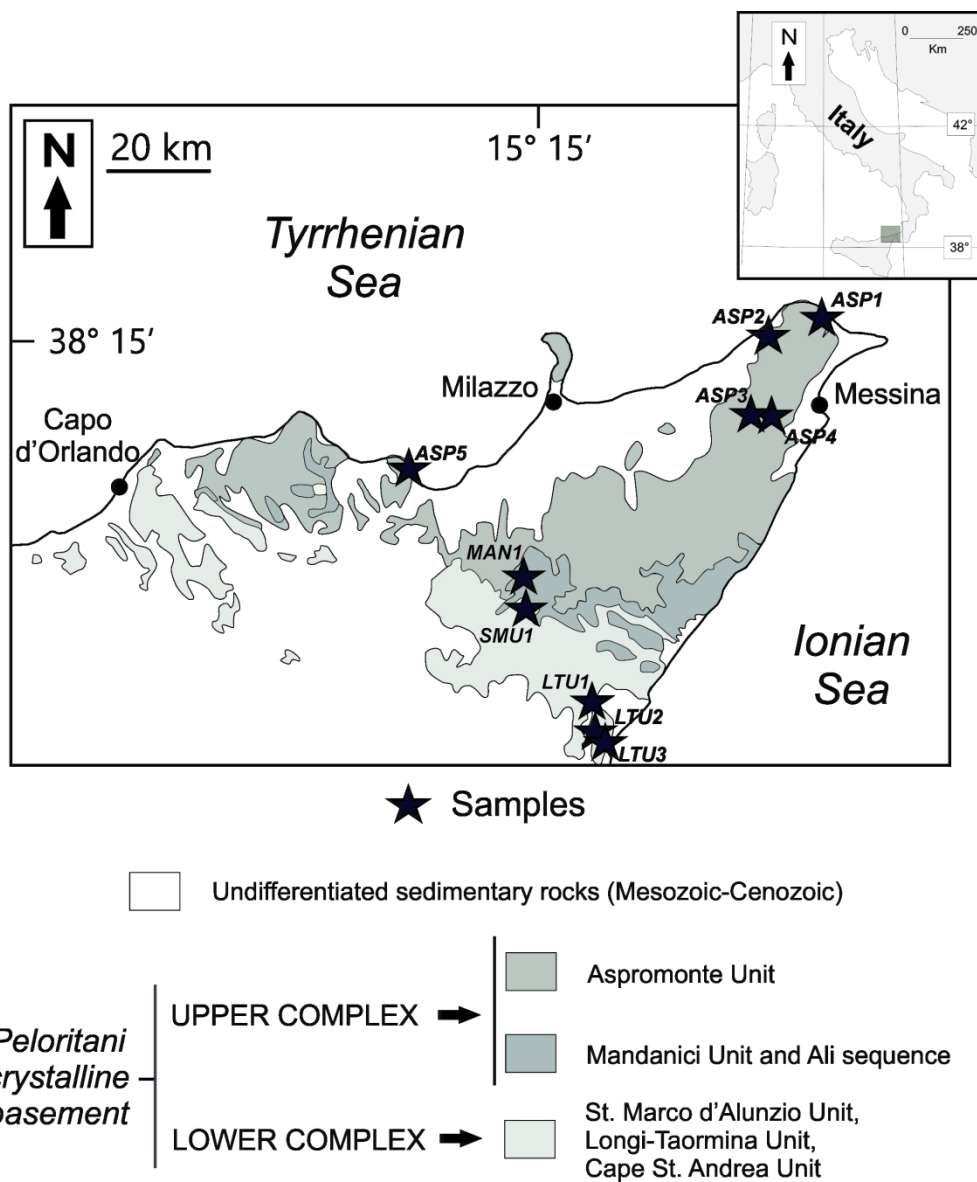


Figure 1. Geological sketch map of the Peloritani Belt (modified from Cirrincione et al.(15)); black stars: location of the collected samples.

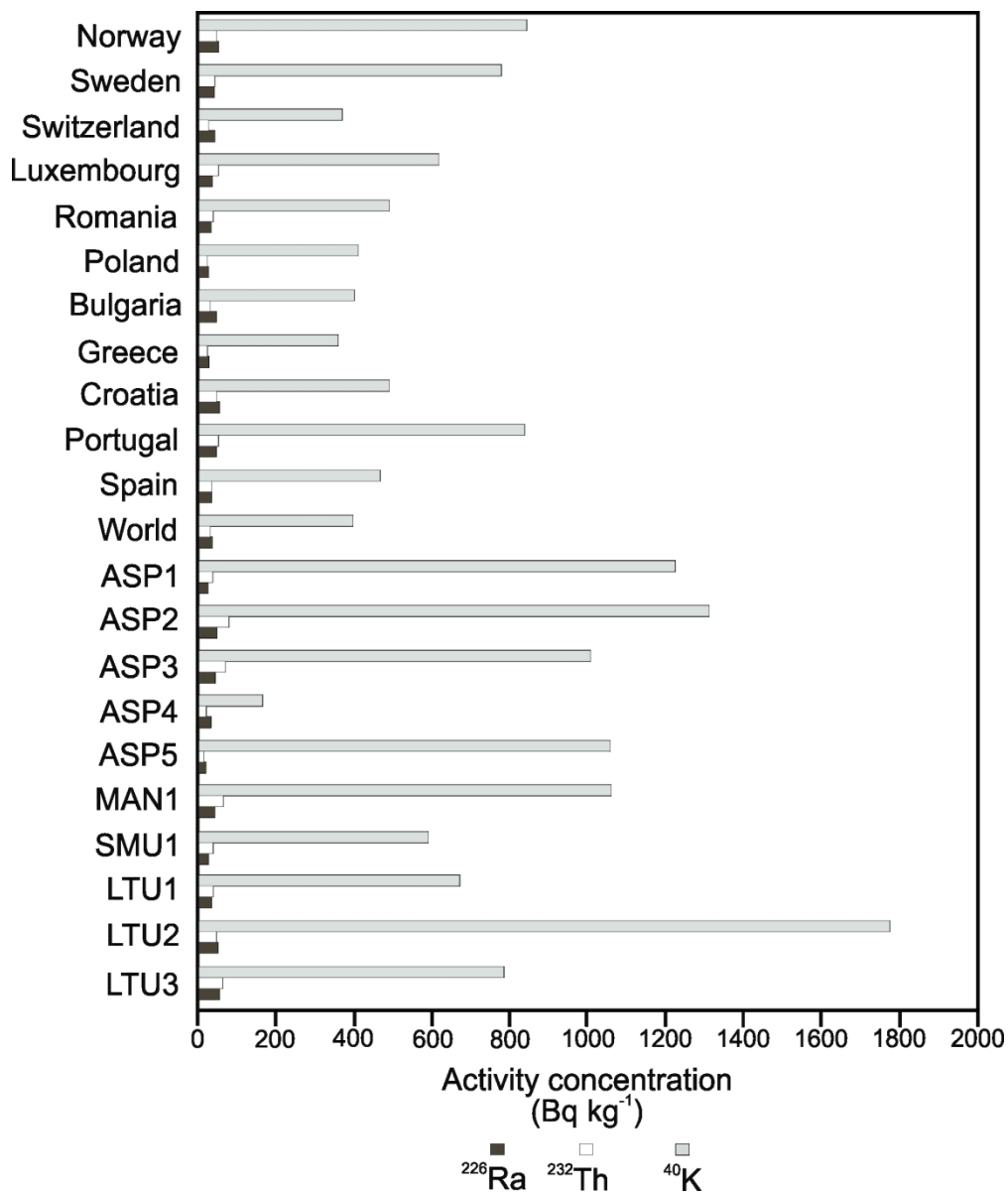


Figure 2. Activity concentrations of ^{226}Ra , ^{232}Th and ^{40}K in the analysed samples compared with those of other European countries and with the worldwide average(1).